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| 1 | A case study on the effects of data temporal resolution on the simulation of water |
|--------|---|
| 2 | flux extremes using a process-based model at the grassland field scale |
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15 Abstract

16 Projected changes to rainfall patterns may exacerbate existing risks posed by flooding. 17 Furthermore, increased surface runoff from agricultural land increases pollution through 18 nutrient losses. Agricultural systems are complex because they are managed in individual 19 fields, and it is impractical to provide resources to monitor their water fluxes. In this respect, 20 modelling provides an inexpensive tool for simulating fluxes. At the field-scale, a daily time-21 step is used routinely, however, it was hypothesized that a finer time-step will provide more 22 accurate identification of peak fluxes. To investigate this, the process-based SPACSYS 23 model that simulates water fluxes, soil carbon and nitrogen cycling as well as plant growth 24 with a daily time-step was adapted to provide sub-daily simulations. As a case study, the 25 water flux simulations were checked against a 15-minute measured water flux dataset from 26 April 2013 to February 2016 from a pasture within a monitored grassland research farm, 27 where the data were up-scaled to hourly, 6-hourly and daily. Analyses were conducted with 28 respect to model performance for: (a) each of the four data resolutions, separately (15-minute 29 measured versus 15-minute simulated; hourly measured versus hourly simulated; etc.); and 30 (b) at the daily resolution only, where 15-minute, hourly and 6-hourly simulations were each 31 aggregated to the daily scale. Comparison between measured and simulated fluxes at the four 32 resolutions revealed that hourly simulations provided the smallest misclassification rate for 33 identifying water flux peaks. Conversely, aggregating to the daily scale using either 15-34 minute or hourly simulations increased accuracy, both in prediction of general trends and 35 identification of peak fluxes. For the latter investigation, the improved identification of 36 extremes resulted in 9 out of 11 peak flow events being correctly identified with only 2 false 37 positives, compared with 5 peaks being identified with 4 false positives of the usual daily 38 simulations. Increased peak flow detection accuracy has the potential to provide clear field 39 management benefits in reducing nutrient losses to water.

40 Key words: SPACSYS; extreme flows; North Wyke Farm Platform; scale effects; grassland;
41

42 **1** Introduction

43 Flooding in the UK puts more than 5 million people in 2.4 million properties at risk each year (Environment Agency, 2009). Projected changes to rainfall patterns (Watts and Anderson, 44 45 2016) may exacerbate the existing risks posed by flooding. Flash flooding or surface water 46 flooding, defined as those flood events where the rise in water is either during or within a few 47 hours of the rainfall that produces the rise, is one of the most common types of flooding in 48 the UK. The utilised agricultural area, of which almost 60% is permanent grassland, covers 49 71% of the total land of the UK (Department for Environment, Food and Rural Affairs, 50 2019). Water fluxes or surface runoff generated from agricultural land can contribute 51 significantly to local floods and nutrient losses that cause water pollution. Flooding of 52 farmland is likely to become more frequent in some areas under projected climate change 53 (Brown et al., 2016), although intriguingly, studies have found increases in precipitation 54 extremes do not necessarily mean increases in flood magnitude, due to decreased soil 55 moisture at storm onset and reduced storm durations (Sharma et al., 2018; Wasko et al., 56 2019). Further, soil erosion is accelerating due to more intense rainfall, leading to the loss of 57 valuable topsoil and the pollution of watercourses (Morison and Matthews, 2016).

58

Accurate forecasting of water runoff (or water fluxes) from agricultural land is, therefore, not only a vital component of flood early-warning systems, but also for associated management strategies for nutrient loss and water pollution. Water fluxes from the soil surface are controlled by soil properties. Long-term hydrological studies have shown that sandy Alfisols can generate higher runoff compared to clayey Vertisols (Pathak et al., 2013), and a greater

risk of flooding on clay soils has been reported (Charlton et al., 2010). The wetness of the soil
before a precipitation event (Merz and Plate, 1997) and soil compaction also affect water
fluxes. Farm machinery and livestock (Adimassu et al., 2019; Alaoui et al., 2018; Newell
Price et al., 2012) can cause serious compaction and so exacerbate flood risk. Natural events,
particularly long and intense precipitation events (Archer and Fowler, 2018), and land cover
variation (Dadson et al., 2017; Keesstra et al., 2018) also affect flux.

70

71 Agricultural systems are complex because they are generally managed at the field scale and 72 each field has its own unique set of soil conditions and topology. Monitoring water surface 73 fluxes in fields is costly both in time and financially. In this respect, modelling provides an 74 effective tool for simulating or forecasting water fluxes. The SPACSYS model (Wu et al., 2007) is one such process-based model. It is a field scale and weather-driven dynamic 75 76 simulation model. Since it was first published in 2007, it has been developed to provide 77 added functionality (Bingham and Wu, 2011; Liu et al., 2013; Wu et al., 2019; Wu et al., 78 2015). The model can simulate the interactions of soil carbon (C), nitrogen (N) and 79 phosphorus (P), plant growth and development, water re-distribution and heat transformation 80 in agricultural fields. The model has been used to investigate several issues including resource use efficiency by crops (Wu et al., 2009), greenhouse gas (GHG) emissions (Abalos 81 82 et al., 2016; Perego et al., 2016), the responses of cropping/grassland systems to environmental change (Wu et al., 2016) and the forecasting of crop yield and stocks of C and 83 84 nutrients (Zhang et al., 2016) under various climatic and soil conditions.

85

86 The SPACSYS model has been developed to investigate not only temporal dynamics, but
87 also within-field spatial variation in processes such as water runoff, using a linked, grid-based

88 approach (grid-to-grid) (Liu et al., 2018). As in all previous implementations of SPACSYS, 89 and common to many agriculture-focused models (Ahuja et al., 2002), a daily time-step was used. However, model predictions of water flux did not increase in accuracy when 90 91 considering grid connectivity. We hypothesise, that a finer time-step might provide this 92 improvement instead; not only in the grid-to-grid model, but also in the (non-grid-to-grid) 93 standard model, as investigated here. Although not demonstrated within this study, increasing 94 the accuracy of water flux simulations should implicitly increase the accuracy of associated SPACSYS simulations, such as those for nutrient loss that use predicted water flux in their 95 96 calculation.

97

98 For our case study, we used measured 15-minute water flux data from one field (or sub-99 catchment) of the North Wyke Farm Platform (NWFP). The NWFP is a systems scale 100 research facility in the south-west of England for investigation of the sustainability of 101 lowland ruminant production systems (Orr et al., 2016). South-west England has a relatively 102 wet climate where the greatest rainfall is in winter and the driest times are between April to 103 July. August tends to show an increase in rainfall over July and starts the inexorable rise in 104 rainfall into autumn and early winter. More recently, the number of flood events has 105 increased (Stevens et al., 2016), mostly in the autumn and winter months; all as a likely 106 consequence of increased surface water runoff (Palmer and Smith, 2013).

107

For this study, the NWFP's 15-minute water flux data were up-scaled to hourly, 6-hourly and daily data and the SPACSYS model was adapted to provide corresponding downscaled simulations at 15-minute, hourly and 6-hourly resolutions (in addition to its usual daily output). This provided four measured water flux datasets and four simulated water flux

datasets over a study period of 34 months (April 2013 to February 2016). Simulations were generated using the same field management practices and parameter configurations. These rich water flux datasets enabled investigation of the effects of temporal scale on model performance not only in terms of extreme water runoff, which is the study focus and provides it's novelty, but also in terms of general trends.

117

118 2 Materials and Methods

119 2.1 Model description

The SPACSYS model includes a plant growth and development component, an N cycling component, a C cycling component, a P cycling component, plus a soil water component that includes representation of water flow to field drains as well as downwards through the soil layers, together with a heat transfer component. The equations to quantify such different processes have been described elsewhere (Liu et al., 2013; Wu et al., 2019; Wu et al., 2007; Wu et al., 2015). Here, only the processes influencing directly the soil water component are presented.

127

128 For SPACSYS, the Richard's equation for water potential and Fourier's equation for 129 temperature are used to simulate water and heat fluxes, which are inherited from the SOIL 130 model (Jansson, 1998). If the water content in a layer rises above a specified value a 131 proportion is held in macropores such that rapid downward water movement takes place due 132 to gravitational forces alone. Water flow from the soil profile to a drainage pipe occurs when the ground water table is above the bottom level of the pipe and the soil below the ground 133 134 water table is saturated. The Hooghoudt drainage flow equation with modification is adopted 135 for the subsurface drainage flow.

136

137 The main processes concerning plant growth in SPACSYS are plant development, 138 assimilation, respiration, root growth and development, water uptake, nutrient uptake, biological N fixation for legume plants and partitioning of photosynthate and nutrients from 139 140 uptake estimated with various mechanisms implemented in the model. N cycling coupled 141 with C cycling covers the transformation processes for organic matter and inorganic N. The 142 main processes and transformations causing size changes to mineral N pools are 143 mineralization, nitrification, denitrification including N gaseous emission and plant N uptake. 144 P cycling is linked to other components such as the plant component, heat transformation and 145 the water cycle. Organic P is subdivided into certain sub-pools with different forms which are 146 connected with transformation rates.

147

148 2.2 The North Wyke Farm Platform

149 The study site is located in south-west England, at the NWFP, Rothamsted Research, Okehampton, Devon (50°46'10''N, 30°54'05''W). For the period 1985-2015, the mean 150 151 annual temperature in North Wyke ranges between 6.8 and 13.4 °C, the mean annual rainfall is 1033 mm and the climate is classed as cool temperate. The platform is a 63 ha systems-152 153 based experimental facility divided into 15 hydrologically isolated sub-catchments across 154 three 21 ha farmlets with five sub-catchments in each. At the time of this study, all three 155 farmlets were used solely for grazing livestock research (sheep and cattle) where each farmlet 156 was operating under different sward management strategies: no re-seeding (permanent 157 pasture); re-seeded monoculture; and re-seeded legume mix. The platform monitors routinely 158 water runoff and water chemistry in each of the 15 sub-catchments, together with other 159 primary data collections (e.g. greenhouse gas emissions, livestock performance) so that each

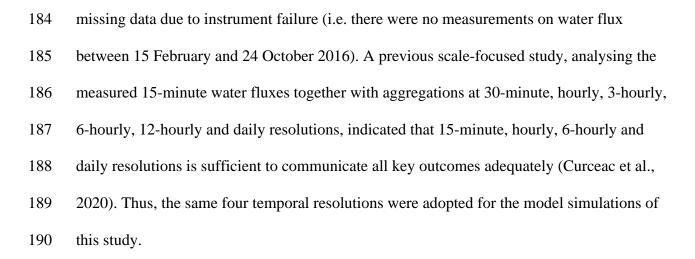
- 160 farming system can be described, contrasted and compared according to its level of
- 161 sustainability (Orr et al., 2016). Datasets are freely available from
- 162 <u>https://www.rothamsted.ac.uk/north-wyke-farm-platform</u>, including those used in this study.
- 163

164 2.3 Model configuration

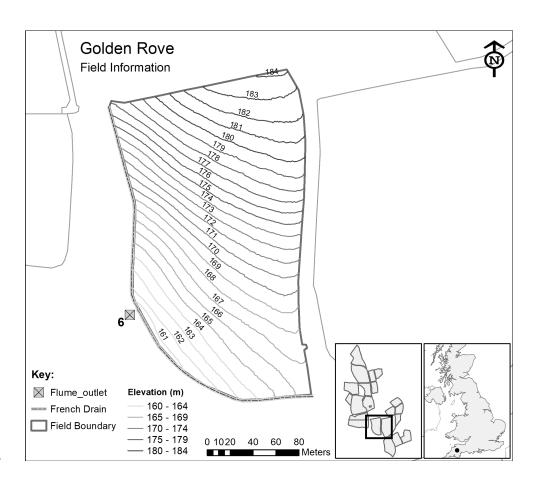
165 For this study, we focused on water fluxes for one sub-catchment in the permanent pasture 166 farmlet called 'Golden Rove'; a single field that has been under permanent pasture since the outset of the platform in 2010 (Fig. Figure 1). The soil class for this field is primarily 167 168 Halstow, which comprises a slightly stony clay loam topsoil (approximately 36% clay) that 169 overlies a mottled stony clay (approximately 60% clay), derived from underlying 170 Carboniferous Culm rocks (Harrod and Hogan, 2008). The study field also has a smaller, but 171 not insignificant area of Denbigh-Cherubeer soil class. In the simulations, the soil type was 172 ignored.

173

174 To mimic the grazing system, daily grass intake and excretion of sheep and cattle in the field 175 was estimated before running the simulations (Carswell et al., 2019; Wu et al., 2016). Soil 176 physical and chemical properties of the field were adopted from a previous study of the same 177 field (Wu et al., 2016). The temporal frequency for the measured water fluxes (1 s^{-1}) from a 178 NWFP water flume has been 15-minutes since the outset of the platform's setup in October 179 2012. However, meteorological measurements at the same 15-minute resolution were only 180 initiated from 30 April 2013. Thus, to ensure consistency in the frequency of the driving variables and the water flux as an output variable, simulations also started from 30 April 181 182 2013. An end-date of 15 February 2016 was chosen to give an interrupted data collection 183 time period of 34 months. A longer time period would entail having significant periods of







192

193 Figure 1. Details of the NWFP sub-catchment selected for this study (sub-catchment number

194 6 of 15, consisting of a single field called Golden Rove).

196 2.4 Statistical analysis

197 Two distinct sets of statistical analyses were conducted with respect to model performance 198 and data resolution: (a) model performance for each of the four data temporal resolutions, 199 separately (i.e. 15-minute measured versus 15-minute simulated; hourly measured versus 200 hourly simulated; 6-hourly measured versus 6-hourly simulated; daily measured versus daily 201 simulated); and (b) model performance conducted at the daily temporal resolution only, 202 where 15-minute, hourly and 6-hourly simulations were each aggregated to the daily scale. 203 The latter analyses provide valuable insights into the worth of using fine temporal resolution 204 data to increase the accuracy of daily simulations, especially with respect to the accurate 205 identification of extremes. This is important as many process-based models in the literature 206 simulate only at a daily time-step (e.g. Del Grosso et al., 2009).

207

208 2.4.1 Model performance graphics

209 Model performance graphics consist of time-series plots, density plots and scatterplots of 210 measured and simulated datasets. For the latter, the ideal 1:1 line, a linear regression fit, and a 211 non-linear regression fit (i.e., a Loess smoother fit; Cleveland, 1979) are given where he 212 estimated intercept and slope parameters from the linear fit should equal zero and one for 213 perfect model simulations, respectively. Results (p-values) from a linear hypothesis test are 214 reported comparing this ideal model with the estimated model using a finite sample F test 215 (see Fox, 2016). The non-linear regression provides added insight into where the simulated 216 values tend to over- or under-predict (e.g., at measured low or high values, respectively). 217 Time-series plots for the errors (i.e. measured minus simulated data) are also given.

219 2.4.2 Model performance indices

220 To further assess the accuracy of the simulations, six accuracy indices were calculated: the

221 mean absolute error (MAE), the normalized root mean square error (NRMSE), the percentage

- bias (PBIAS), the Nash-Sutcliffe efficiency (NSE), the index of agreement (d) and the Kling-
- 223 Gupta efficiency (KGE), as given in Table 1.
- 224
- 225 Table 1. Accuracy indices formulae, where \hat{z}_i are the simulated values, z_i are the measured

226 values, \overline{z}_i is the mean of the measured values, r is the Pearson product-moment correlation

227 coefficient (between measured and simulated) and σ is the standard deviation.

| Index form | Index formula | Min. | Max. | Ideal |
|------------|---|------|------|-------|
| Error | $MAE = \frac{1}{N} \sum_{i=1}^{N} \hat{z}_i - z_i $ | 0 | œ | 0 |
| Error | NRMSE = $100 \frac{\sqrt{\frac{1}{N} \sum_{i=1}^{N} (\hat{z}_i - z_i)^2}}{z_{max} - z_{min}}$ | 0 | 8 | 0 |
| Error | $PBIAS = 100 \frac{\sum_{i=1}^{N} (\hat{z}_i - z_i)}{\sum_{i=1}^{N} z_i}$ | 0 | 8 | 0 |
| Agreement | NSE = $1 - \frac{\sum_{i=1}^{N} (\hat{z}_i - z_i)^2}{\sum_{i=1}^{N} (z_i - \overline{z}_i)^2}$ | -∞ | 1 | 1 |
| Agreement | $d = 1 - \frac{\sum_{i=1}^{N} (\hat{z}_i - z_i)^2}{\sum_{i=1}^{N} (\hat{z}_i - \overline{z}_i + z_i - \overline{z}_i)^2}$ | 0 | 1 | 1 |

Agreement
$$\mathsf{KGE} = 1 - \sqrt{(\mathsf{r} - 1)^2 + \left(\frac{\sigma_{\hat{z}}}{\sigma_z} - 1\right)^2 + \left(\frac{\overline{z}}{\overline{z}} - 1\right)^2} \quad -\infty \quad 1 \quad 1$$

228

229 2.4.3 Simulation accuracy of measured peaks

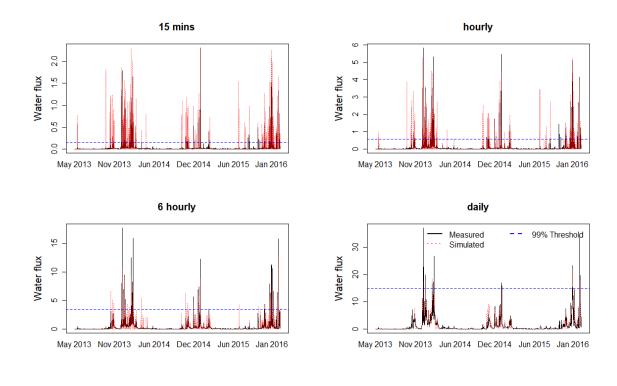
230 To investigate model accuracy in simulating water flux peaks, a threshold at the 99th 231 percentile of each measured water flux dataset was used to identify peak flows. Model 232 simulations were then assessed to determine if they could similarly exceed this threshold 233 coinciding with a measured exceedance. Incidences of correct peak flow simulations, false 234 negatives (simulation does not exceed threshold when measured flow does), false positives 235 (simulation exceeds threshold when measured flow does not) and corresponding Kappa 236 values are reported. The Kappa statistic provides a measure of agreement beyond the level of 237 agreement expected by chance alone. General guidelines for Kappa values are as follows: less 238 than 0.2 slight agreement, 0.2 to 0.4 fair agreement, 0.4 to 0.6 moderate agreement, 0.6 to 0.8 239 substantial agreement, greater than 0.8 almost perfect agreement, and equal to 1 perfect 240 agreement.

241

242 **3 Results**

243 3.1 Model performance for each of the four data temporal resolutions, separately

Comparisons between the measured and simulated water flux rates at different temporal resolutions are shown in Fig. Figure 2. Visually, it appears that simulations of daily and 6hourly water fluxes tend to under-predict the measured data, often missing high peaks, while simulations of 15-minute and hourly data possibly tend to over-predict. However, the scatterplots of the measured and simulated data, together with the ideal 1:1 line, a linear regression fit, and a Loess smoother fit (Fig. Figure 3) present a more complete picture.
Simulations for all four temporal resolutions clearly tend to over-predict, with the level of
over-prediction increasing as the resolution increases. Over-prediction is shown with each
linear regression fit lying below the 1:1 line; and increasingly so, as the resolution increases.
All linear regression fits were found to be significantly different to the 1:1 line, each with *F*test *p*-values < 0.0001.





256

Figure 2. Time-series plots for measured and simulated water flux data (not aggregated) for 15-minute, hourly, 6-hourly and daily data (in units of mm 15min⁻¹, mm h⁻¹, mm 6h⁻¹ and mm d⁻¹, respectively). All plots are shown with a threshold at the 99th percentile of measured data (at 0.138 mm 15min⁻¹, 0.553 mm h⁻¹, 3.45 mm 6h⁻¹ and 14.9 mm d⁻¹, respectively).

261

262 For all four temporal resolutions, the tendency to over-predict decreases at the largest

263 measured water fluxes, as shown by the concave behaviour of the loess smoother fit (Fig.

264 Figure 3), with daily simulations tending to under-predict at very large fluxes, thus, missing 265 extreme events that may cause flooding and associated nutrient and sediment losses. Clearly, 266 'smoothing bias' increases as temporal resolution decreases. The 15-minute simulations 267 maintain the variation shown in the measured data (i.e. observations range from 0 to 2.306 mm 15min⁻¹ while simulations range from 0 to 2.310 mm 15min⁻¹), while the daily 268 simulations do not (i.e. observations range from 0 to 36.97 mm d⁻¹ while simulations only 269 270 range from 0 to 22.20 mm d⁻¹). As each 'simulation-to-observation' comparison is on a 271 different scale, it is not useful to present further model fit diagnostics, such as error and 272 agreement indices.

273

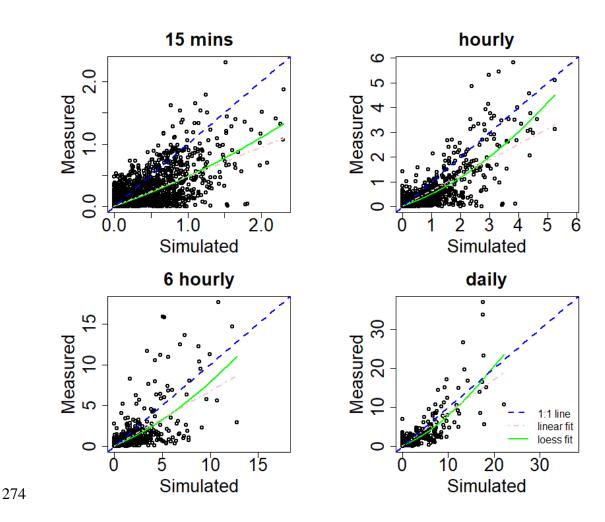


Figure 3. Scatterplots of the measured and simulated data (not aggregated) for 15-minute,

hourly, 6-hourly and daily data. Scatterplots are shown with the 1:1 line, a linear regression

fit and a loess smoother fit. Units are in mm 15min⁻¹, mm h⁻¹, mm 6h⁻¹ and mm d⁻¹,

278 respectively.

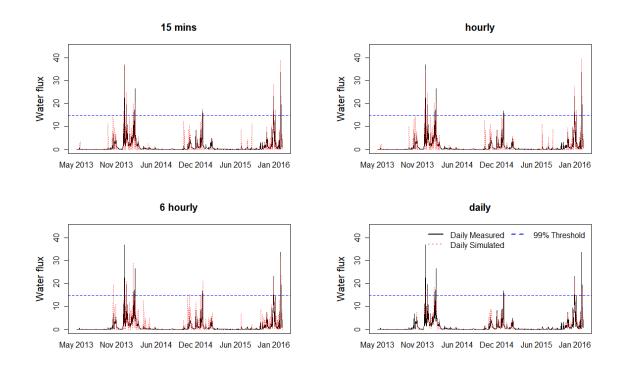
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280 3.2 Model performance for simulations aggregated to the daily scale

281 Comparisons between the measured and simulated water flux rates aggregated to the same daily scale are shown in Fig. Figure 4. There are clear instances of both over- and under-282 283 prediction for all four daily outputs. The scatterplots (Fig. Figure 5) of daily measured and 284 daily simulated data from different aggregations, together with the 1:1 line, a linear fit, and a 285 loess smoother fit, again provide a clear visualisation of the relations in the time-series plots. 286 Simulations for all three aggregations to daily (15-minute, hourly and 6-hourly) again tend to 287 over-predict (as their linear fits lie below the 1:1 line), but this over-prediction is broadly 288 similar across the four datasets, and not as great as that found with the unaggregated data, 289 above. The 6-hourly aggregations appear to be the least accurate. Again, all linear regression 290 fits were found to be significantly different to the 1:1 line, each with *F*-test *p*-values < 0.0001. 291

292

In this instance, 'smoothing boas' increases as aggregation resolution decreases, where simulations for 15-minute and hourly aggregations both increase the variation shown in the measured daily data (i.e. 0 to 36.97 mm d⁻¹); with 15-minute daily aggregations ranging from 0 to 38.94 mm d⁻¹ and hourly daily aggregations ranging from 0 to 39.64 mm d⁻¹. Conversely, the 6-hourly daily aggregations and the daily simulations reduce variation with the 6-hourly daily aggregations ranging from 0 to 31.70 mm d⁻¹ and the (unaggregated) daily simulations ranging from 0 to 22.20 mm d⁻¹. In summary, daily simulations based on component 15minute and hourly aggregations have the potential to identify peak water fluxes (and, thus,
flood events) and predict their magnitudes more accurately, relative to 6-hourly aggregations
and (unaggregated) daily simulations.



303

Figure 4. Time-series plots for daily measured and daily simulated water flux data (with the first three plots having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily). All units in mm d⁻¹. All plots are shown with a threshold at the 99th percentile of measured data (14.90 mm d⁻¹).

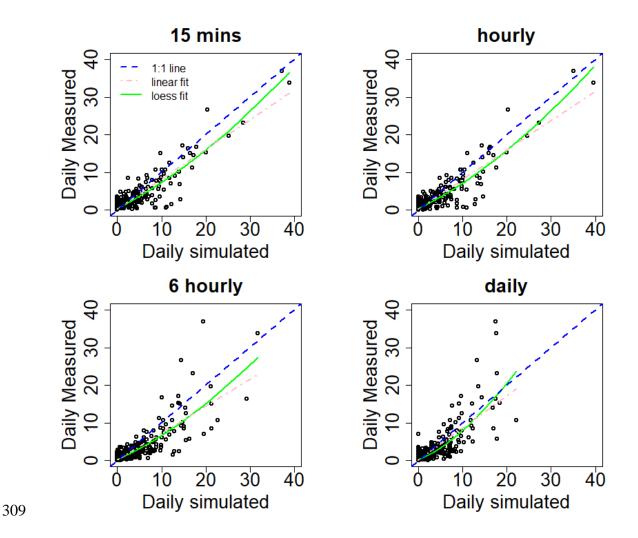
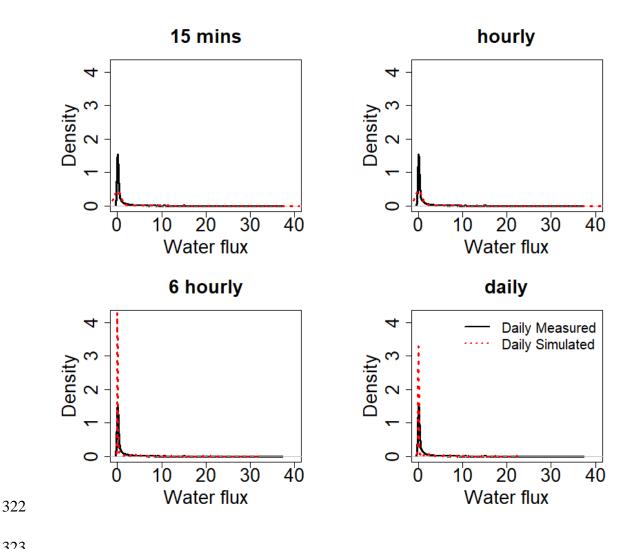


Figure 5. Scatterplots of the daily measured and daily simulated water flux (with the first three plots having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily). Scatterplots are shown with the ideal 1:1 line, a linear regression fit and a loess smoother fit. All units in mm d⁻¹.

Further clarity on bias is provided in the density plots for the measured and simulated data (Fig. Figure 6). Here, daily simulations based on 15-minute and hourly aggregations have a lower density at small daily water fluxes than that found with the measured data, while the 6hourly aggregations and (unaggregated) daily simulations have a higher density at small daily water fluxes. This is combined with a longer tail in the density curve for the 15-minute and

320 hourly aggregations, as each can simulate large daily water fluxes, while the 6-hourly 321 aggregation and (unaggregated) daily simulations do not have this property.



323

324 Figure 6. Density plots for daily measured and daily simulated data (with the first three plots 325 having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily). All units 326 in mm d^{-1} .

328 The error indices (MAE, NRMSE and PBIAS) are reported for each daily aggregation in Fig. 329 Figure 7, where the 15-minute and hourly aggregations clearly perform more accurately than 330 the 6-hourly aggregation and (unaggregated) daily simulations. Errors (i.e. residuals) are also

reported over the study time period in Fig. Figure 8, where errors tend to be larger for the
daily simulations based on the 6-hourly aggregation and the (unaggregated) daily simulations.
Interestingly, the 6-hourly aggregation consistently is the least accurate, including being less
accurate than the (unaggregated) daily simulations.



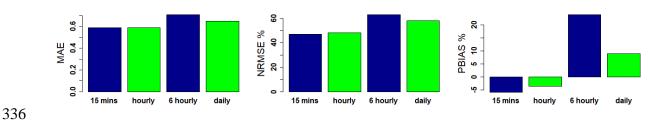


Figure 7. Error (MAE, NRMSE, PBIAS) indices with respect to daily measured and daily
simulated data (with 15-minute, hourly, 6-hourly data aggregated to daily).

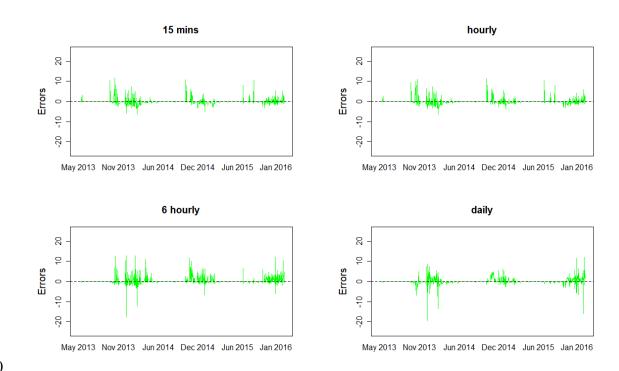




Figure 8. Time-series of errors (simulated minus measured data) aggregated to the daily
temporal resolution. All units in mm d⁻¹. Positive errors represent over-prediction by the
model.

344

Agreement indices (NSE, d and KGE) are reported for each daily aggregation in Fig. Figure 9, where again the 15-minute and hourly aggregations perform more accurately than the 6hourly aggregation and (unaggregated) daily simulations (although daily simulations perform relatively well according to the KGE index). From the given accuracy diagnostics, it is not immediately apparent whether daily simulations based on 15-minute or hourly aggregations are the most accurate, and as such, both appear to increase the accuracy relative to SPACSYS' daily simulations. Again, the 6-hourly aggregation is the least accurate.

352

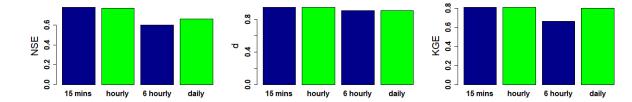




Figure 9. Agreement (NSE, *d*, KGE) indices with respect to daily measured and daily
simulated data (with 15-minute, hourly, 6-hourly simulations aggregated to daily).

356

357 3.3 Simulation of measured peaks

358 To investigate the ability of the model to simulate and identify water flux peaks, the 99th

359 percentile of each measured water flux dataset was used as a threshold to identify peak flows,

as highlighted in Figs. Figure 2 and Figure 4 (the dashed blue line). Incidences of correct

361 peak flow simulations, false negatives, false positives and the resultant Kappa values are 362 given in Table 2. It appears that hourly simulations provide the largest correct classification 363 rate (Kappa = 0.553) for the unaggregated approach, but with only moderate success in 364 identifying measured peak flows (as a promising 92% identification rate is tempered by a poor mis-identification rate). Conversely, aggregating to the daily scale using either 15-365 366 minute or hourly simulations was able to provide much greater agreement in identifying 367 measured peak flows at the daily scale, with each identifying 9 out of 11 peak flow events 368 correctly, coupled with only 2 false positives (Kappa = 0.816 in both cases). This level of 369 agreement was far greater than that found through directly simulating the daily data, which 370 provided only moderate success in identifying measured peak flows (Kappa = 0.495). Again, 371 the 6-hourly aggregation is the least accurate with a relatively high number of false positives 372 (simulated flow exceeds the threshold when measured flow does not).

Table 2. Accuracy at peak water fluxes according to simulation resolution. Peaks taken at 99th
percentile of measured data (see the dashed blue line in Figs. Figure 2 and Figure 4).

| Simulation | Sample | Measured | Correctly | False | False | Kappa |
|---------------------|--------|----------|-----------|----------|----------|-------|
| resolution | size | Peaks | Simulated | Negative | Positive | |
| Unaggregated | | | | | | |
| 15-minute | 97920 | 980 | 759 (77)* | 221 | 1224 | 0.506 |
| hourly | 24480 | 245 | 225 (92) | 20 | 335 | 0.553 |
| 6-hourly | 4080 | 41 | 32 (78) | 9 | 52 | 0.503 |
| daily | 1020 | 11 | 5 (45) | 6 | 4 | 0.495 |
| Aggregated to daily | | | | | | |
| 15-minute | 1020 | 11 | 9 (82) | 2 | 2 | 0.816 |

| hourly | 1020 | 11 | 9 (82) | 2 | 2 | 0.816 |
|----------|------|----|--------|---|---|-------|
| 6-hourly | 1020 | 11 | 6 (55) | 5 | 9 | 0.455 |
| daily | 1020 | 11 | 5 (45) | 6 | 4 | 0.495 |

^{*} Value in brackets shows a percentage of correctly simulated peaks to measured peaks.

377 **4 Discussion**

378 4.1 Model performance

379 4.1.1 Unaggregated data

380 The statistical analyses for model performance suggested that the SPACSYS model simulates 381 the general trend of water fluxes at the four different temporal resolutions reasonably well 382 (Figs. Figure 2 and Figure 3). All simulations tended to over-predict water flux, however, and 383 only simulations at the finest resolutions maintained the variation in the measured data. The 384 accuracy of water flux peak simulations varied among the four resolutions (Table 2). Almost 385 92% of the measured peaks over the simulated period were modelled correctly at an hourly 386 resolution, the resolution with the smallest misclassification rate. However, this was tempered 387 by a high rate of predicting peaks that did not exist. A previous statistical analysis of peak 388 flows at different scales from a different NWFP sub-catchment (similar in size to the one 389 used here), modelled and simulated by a Generalized Pareto distribution, also showed the 390 greatest agreement at the hourly resolution (Curceac et al., 2020).

391

392 4.1.2 Aggregated to Daily

When simulations at a finer temporal resolution were aggregated to a daily rate, thesimulation results using both the 15-minute and hourly aggregations showed the greatest

accuracy broadly equally, both in the prediction of general trends (Figs. Figure 4 to Figure 9)

396 and the identification of peak flows (Table 2). This demonstrates clearly that the daily 397 simulation of water fluxes with the SPACSYS model, informed by finer temporal resolution 398 data, can increase simulation accuracy. This result is an important advance relative to 399 previous SPACSYS studies, which only used a daily time-step, and which similarly used sub-400 catchments of the NWFP as the study site (Liu et al., 2018). 401 The simulation results using both the 15-minute and hourly aggregations showed the greatest 402 accuracy broadly equally. When the simulation time step is getting longer, average 403 precipitation intensity might be weaker, which causes simulated water fluxes smoother and 404 the deviations with monitored water fluxes larger. Given the complexity of the soil-water 405 processes that operate across the field, it is not surprising to see substantial variation around 406 the 1-1 line when predicting flow at 15-minute to 6-hour scales due to inherent variability. 407 We would expect to see a similar phenomenon with any similar process-based model. What 408 is important is that when these fine-scale predictions are aggregated we get substantial 409 improvements to daily predictions.

410 4.2 Results in context and their generalisation

411 Results are consistent with other studies that similarly showed that differences in the 412 (unaggregated or aggregated) time-step have the greatest impact on runoff simulation 413 accuracy relative to other factors, some of which, also investigated changes in spatial 414 resolution (i.e. aggregating over different spatial units) (Choi et al., 2018; Huang et al., 2019; 415 Jeong et al., 2010; Kavetski et al., 2011; Merz et al., 2009). Thus, the value of using 416 aggregated fine temporal resolution simulations to increase the accuracy of daily simulations 417 can be said to hold generally for other process-based models provided the hydrological 418 process is described appropriately. However, it does not follow that daily simulation accuracy 419 will continue to increase as the temporal resolution of the aggregated data becomes finer. 420 This study found aggregating hourly simulations to daily to be just as accurate as aggregating

421 15-minute simulations to daily, while aggregating 6-hourly simulations to daily performed422 less well than the usual daily simulations.

423

424 The temporal resolution for process-based models should be chosen carefully to balance 425 between capturing all important processes, the study objectives and data availability. For our 426 study, with flooding as context, the identification of water flux extremes in a grassland field 427 (or small sub-catchment) with a heavy clayed soil, is viewed as the important process, more 428 so than capturing broad trends in water flux. It is well-known that running a model at a finer 429 resolution, then aggregate, will increase the prediction accuracy in a broad sense (see above). 430 What has received less attention in the literature is the effects of temporal resolution on a 431 model's ability to capture extremes (e.g. see Schaller et al., 2020, in the context of 432 streamflow). In this respect, our study has found daily peak flows to be more accurately 433 identified using aggregations of simulations at finer resolutions, than using coarse daily 434 simulations directly. Of course, measurement at a finer resolution comes at a cost and this 435 needs to be balanced with associated improvements in model accuracy. In this instance, this 436 interplay is simple to resolve since aggregating to the daily scale using both 15-minute and 437 hourly simulations were equally as accurate, meaning measuring at an hourly interval would 438 be sufficient for the case study site.

439

The appropriate temporal resolution to simulate water fluxes using a hydrological model
depends on hydro-climatological and geophysical characteristics, and the scale of the process.
It has been suggested that an appropriate temporal resolution could be between 12 hours for
middle-sized upstream areas and 48 hours for a complete river basin (Booij and Tran, 2005).
As the size of the field for this study is < 4 ha, the indicated hourly resolution appears

445 reasonable. Observed and projected changes in the UK's climate suggest an increase in heavy 446 rain events and wetter winters (Committee on Climate Change, 2017), where some UK 447 regions will be more affected than others. This will inevitably change agricultural 448 management practice and land use across the UK. Taking as an example the grazed pasture of 449 this study, introducing a deep-rooting grass suited to its heavy clay soils (Macleod et al., 450 2013) and/or the mechanical loosening of topsoil (Newell-Price et al., 2011) would reduce 451 water runoff, whereas conversion to an arable crop (e.g. wheat) would provide its own set of 452 water runoff influences. Such changes would alter the characteristics of the water fluxes 453 generated, as the field's soil properties will change, meaning the determination of an 454 appropriate resolution to simulate water fluxes may also change from the hourly resolution 455 suggested here. This is analogous to other hydrological studies where, for example, different 456 overflow designs in roof drainage structures have markedly variable responses to rainfall 457 intensity increases (Verstraten et al., 2019).

458

459 4.3 Inputs that impact hydrological model performance

460 Key model input variables such as precipitation can determine the impacts of simulation 461 time-steps on the performance of hydrological models; for example, the duration and 462 temporal variability of a precipitation event in relation to the rainfall-runoff lag time (Ficch) 463 et al., 2016). A multiple-day precipitation event is the main cause of continuous runoff events 464 and related peaks. For example, for this study, there was almost an unbroken measured 465 precipitation period from 14 December 2013 to 5 March 2014, which brought a total of 541 466 mm of precipitation, 78% of which was measured as surface runoff (i.e., measured water 467 flux). Study simulations showed 70, 70, 81 and 85% as water fluxes over the period at the 15-468 minute, hourly, 6-hourly and daily resolutions, respectively. Previous studies showed that 469 wetter soils had less capacity to store water, resulting in greater runoff volumes (Huang et al.,

470 2017; Kibet et al., 2014; Zehe et al., 2010). Both observations and simulations in this study 471 confirmed this finding. Conversely, for a single day event, a measured 92% of 40.2 mm daily 472 precipitation was discharged on 23 December 2013. The simulations generated 99, 99, 90 and 473 44% water losses at the 15-minute, hourly, 6-hourly and daily resolutions, respectively. Thus, 474 almost all of the precipitation contributed to the water loss on this day, where only the daily-475 scale simulation did not capture this. However, although heavy rainfall is necessary to 476 generate water fluxes, it is not a sufficient condition for a higher surface runoff rate to occur 477 (Ledingham et al., 2019). For example, there was about 25 mm precipitation on 14 May 2013 478 and on 13 August 2015, but both the simulations and the observations (at all four resolutions) 479 did not show apparent water fluxes. Further, a daily precipitation of 4 mm on 27 February 480 2015 generated a measured 120% water loss, together with simulated values of 48, 148, 125 481 and 75% water loss at the 15-minute, hourly, 6-hourly and daily resolutions, respectively.

482

483 The generation of water fluxes not only depends on the intensity of precipitation, but also 484 surface coverage, topology and soil physical properties of the field. In hydrology, lag time, 485 defined as the time difference between the peak runoff and mass centre of rainfall excess 486 (Hall, 1984), is usually used to determine a runoff rate. Although the SPACSYS model does 487 not use this parameter to estimate the surface runoff rate, it uses the Richard's equation to 488 calculate soil water redistribution where soil hydraulic conductivity, saturated water content 489 and plant uptake play critical roles in water infiltration and consequently surface runoff. A 490 trial study on the spatial variation of soil hydraulic conductivity (unpublished data) in a 491 NWFP field, nearby to the study field, highlighted clear within-field variation, partially 492 because of compaction caused by grazed animal movement. However, the soil physical 493 properties used in the simulations were estimated based on soil texture, and at the field-scale 494 only. The error and uncertainty introduced by this approach are likely to be transferred to the 495 errors in simulating infiltration and surface runoff rates. To improve model simulation
496 accuracy, soil physical properties as core information should be provided wherever possible.

497

498 4.4 Further considerations of scale

499 The processes controlling water fluxes operate across a range of spatial and temporal scales, 500 and the time-series that are recorded represent an aggregation of these effects. For example, 501 effects of evapotranspiration will dominate at annual scales whereas more local impacts of 502 precipitation will manifest at finer scales (Rust et al., 2014). As noted above, the 503 characteristics of the study area, (e.g. size, soil condition and topography) will impact the 504 dominant scales of variation and hence the frequency at which it is most appropriate to model 505 or measure water fluxes. Rust et al. (2014) presented an analysis which aimed to determine 506 whether the process-based model they studied captured the scale-dependent variation 507 measured in catchment runoff. They analysed model residuals using wavelet-based signal 508 processing methods and found that although their model captured broadly the scale-509 dependent variation in the data, fine scale variation was always under-predicted. Our results 510 shed light on their observation as we confirm that if the scale of the model prediction is not 511 sufficiently fine then model-damping will result in an under-prediction of extreme events.

512

513 **5 Conclusions**

For the grassland study site, the adapted process-based model (SPACSYS) could adequately simulate the trends in measured water fluxes and identify their extremes. At a daily time-step, model accuracy increased when simulations were run at finer temporal resolutions, specifically 15-minute and hourly, and then aggregated to daily (a coarse output resolution

commonly used in field-scale agricultural settings). Aggregating using 6-hourly simulations was less accurate. For the study site, which constitutes a field of a grassland research farm platform (NWFP), simulation of water fluxes at an hourly resolution is likely optimal since use of the 15-minute resolution did not increase prediction accuracy or the ability to identify extremes in flow further. Therefore, for modelling purposes, monitoring frequency could be reduced safely to hourly from the current 15-minute resolution.

524

525 Results provide information not only for the NWFP experiment, but also and indirectly, the 526 UK grassland farming regions that its outputs upscale to (Pulley and Collins, 2019). Study 527 results are crucial in relation to meeting the increasing demand for reliable simulation-based 528 runoff forecasts at daily and sub-daily resolutions, where accurate knowledge of peak 529 discharge and stage are essential not only for flood protection, but also to help increase the 530 forecast accuracy of associated emissions such as nutrient or sediment loss, that each use 531 water flux as a component. Further research is called for in specifying the temporal resolution 532 amongst the wide range of field-scale hydrological/agricultural models currently applied. 533 This needs to be coupled with linked changes in climate and land use to increase model 534 forecast accuracy and to optimise data acquisition schemes on farms generally.

535

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721 **Figure captions**

| 722 | Figure 1. Details of the NWFP catchment selected for this study (catchment number 6 of 15, |
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| 723 | consisting of a single field called Golden Rove). |

- Figure 2. Time-series plots for measured and simulated water flux data (not aggregated) for
- 72515-minute, hourly, 6-hourly and daily data (in units of mm 15min⁻¹, mm h⁻¹, mm 6h⁻¹726and mm d⁻¹, respectively). All plots are shown with a threshold at the 99th percentile727of measured data (at 0.138 mm 15min⁻¹, 0.553 mm h⁻¹, 3.45 mm 6h⁻¹ and 14.9 mm d⁻¹,728respectively).
- Figure 3. Scatterplots of the measured and simulated data (not aggregated) for 15-minute,
- hourly, 6-hourly and daily data. Scatterplots are shown with the 1:1 line, a linear
 regression fit and a loess smoother fit. Units are in mm 15min⁻¹, mm h⁻¹, mm 6h⁻¹ and
 mm d⁻¹, respectively.
- Figure 4. Time-series plots for daily measured and daily simulated water flux data (with the first three plots having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily). All units in mm d⁻¹. All plots are shown with a threshold at the 99th percentile of measured data (14.90 mm d⁻¹).
- Figure 5. Scatterplots of the daily measured and daily simulated data (with the first three plots
 having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily).

| 739 | Scatterplots are shown with the ideal 1:1 line, a linear regression fit and a loess |
|-----|---|
| 740 | smoother fit. All units in mm d ⁻¹ . |
| 741 | Figure 6. Density plots for daily measured and daily simulated data (with the first three plots |
| 742 | having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily). |
| 743 | All units in mm d ⁻¹ . |
| 744 | Figure 7. Error (MAE, NRMSE, PBIAS) indices with respect to daily measured and daily |
| 745 | simulated data (with 15-minute, hourly, 6-hourly data aggregated to daily). |
| 746 | Figure 8. Time-series of errors (simulated minus measured data) aggregated to the daily |
| 747 | temporal resolution. All units in mm d ⁻¹ . Positive errors represent over-prediction by |
| 748 | the model. |
| 749 | Figure 9. Agreement (NSE, d, KGE) indices with respect to daily measured and daily |
| 750 | simulated data (with 15-minute, hourly, 6-hourly simulations aggregated to daily). |
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