

1 **Use of untargeted metabolomics to analyse changes in extractable soil organic**
2 **matter in response to long-term fertilization**

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24 **Abstract**

25 This study aimed to explore the soil metabolic response to long-term fertilizer
26 application and the effect of this response on the microbial community, taking
27 advantage of the Woburn Organic Manuring Experiment (UK), which has been running
28 since 1964. Untargeted metabolomes detected by gas chromatography-time of flight
29 mass spectrometer/mass spectrometry (GC-TOFMS/MS) and ultra-high-pressure
30 liquid chromatography-quadrupole time of flight mass-spectrometer/mass
31 spectrometry (UHPLC-QTOFMS/MS) were used to explore which method better
32 reflected soil microbe-accessible metabolites. The microbial community abundance
33 was detected by high-throughput sequencing. We found that long-term farmyard
34 manure application enhanced the total and dissolved C and N contents in the soil. The
35 metabolite content detected by GC-TOFMS/MS (TOF detector with a cold injection
36 unit) had a linear negative correlation to soil organic matter, extractable organic
37 nitrogen (N), and microbial carbon (C). Conversely, the metabolite content detected by
38 UHPLC-QTOFMS/MS was positively correlated to them, indicating the metabolites
39 detected by UHPLC-QTOFMS/MS is the main components of soil soluble organic
40 matter. More positive than negative correlations were observed between metabolites
41 and bacterial (69.5%) or fungal (67.9%) taxa in the co-occurrence network. Among the
42 bacterial taxa in the network, the family Planococcaceae and genus *Paenibacillus*
43 showed the most correlations with metabolites. Extraction and detection methods are
44 affected by not only the variety but also the number of detected metabolites. Careful

45 consideration is needed when selecting which methods to use. We demonstrated a
46 strong correlation between soil metabolites and the microbial community abundance.
47 However, a deeper understanding of soil microbial function and the formation, content,
48 and decomposition of metabolites is still needed.

49 **Keywords:** soil organic matter, dissolved organic matter, chemical fertilizer, farmyard
50 manure, untargeted metabolomes

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53 **Introduction**

54 Most of the C in the terrestrial biosphere is retained as soil organic matter (SOM), which
55 originates from microbes, plants, and animals (Johnston et al. 2004). Dissolved organic
56 matter (DOM) is the most biologically-accessible component of SOM, playing a crucial
57 role in C, N, and sulfur (S) cycling (Ma et al. 2020b, 2021a; Swenson et al. 2015). Soil
58 microorganisms derive metabolites predominantly from SOM and its biomass turnover
59 (Brown et al. 2021; Liang et al. 2019). DOM contains a series of organic matter
60 compounds such as carbohydrates, amino acids, hydroxyl acids, sugar acids,
61 nucleosides, sterols, aromatics, amines, and miscellaneous compounds (Brown et al.
62 2021). DOM is in a constant state of flux driven by the microbial community and *in*
63 *situ* metabolic activities (Mcleod et al. 2021; Schmidt et al. 2011). Therefore,
64 understanding the composition and turnover of soil microbe-accessible substrates is
65 crucial for exploring the complex dynamics of microbial communities and their nutrient
66 cycling (Ma et al. 2020c; Zhu et al. 2022).

67 Fertilization is one of the most important field management interventions that
68 strongly affects soil element content, nutrient cycling, and the microbial community.
69 Annually, agricultural production produces about seven billion tons of farmyard
70 manure (FYM) globally (Thangarajan et al. 2013). Manure application can increase soil
71 structural stability and nutrient levels, enhancing soil C sequestration and biological
72 activity in arable land (Maillard and Angers, 2014). Partly substituting inorganic
73 fertilizer with FYM can sustain agricultural productivity and reduce environmental

74 pollution (Hoyle and Fang, 2018). FYM application strongly stimulates belowground
75 biogeochemical processes: directly by adding large amounts of organic C and nutrients,
76 and indirectly by modifying biotic activity (Liu et al. 2020; Ma et al. 2018). Subsoil
77 differs from topsoil in nutrient content, microbial biomass, community composition,
78 bioavailability, age, and accessibility of soil C, which affect the rates of SOM
79 decomposition (Cheng et al. 2017). In contrast to chemical fertilizers, long-term FYM
80 application generally improves the total and DOM content of topsoil and subsoil (Ma
81 et al. 2020b; Yan et al. 2018). Additionally, it enhances the activities of enzymes such
82 as β -glucosidase, protease, urease, and cellulase (Chang et al. 2010; Ma et al. 2020b).
83 However, how FYM and chemical fertilizer application influences the soil metabolite
84 composition is unclear.

85 A healthy and well-functioning soil system is vital for providing ecosystem
86 services, especially food production in agricultural ecosystems (Liu et al. 2022; Wei et
87 al. 2021). Metabolites in DOM are intermediates or products of enzymatic reactions,
88 including organic acids, sugars, amino acids, and fatty acids. These are involved in
89 microbial growth, development, and function. In addition to molecular methods of soil
90 biological quality assessment, extracting and quantifying primary metabolites offer an
91 alternative approach to better understanding belowground functions. The metabolic
92 approach has been used extensively in plant biology (Hartman et al. 2020), biomedical
93 science (Gupta et al. 2018), and research in the biochemical response of a microbial
94 species (Jozefczuk et al. 2014). However, its application in soil is limited, especially
95 under field conditions, and most studies have only focused on specific metabolites (Ma

96 et al. 2021a; Warren, 2020). Recent studies have shown that the soil metabolome is
97 sensitive and can reflect soil microbial functional responses to changes in their
98 environment, such as fertilizer application, extreme drought, and dry-wet or freeze-
99 thaw events (Brown et al. 2021; Miura et al. 2020).

100 Traditionally, soil DOM quantification is achieved by extraction from soil
101 samples using specific solutions (water, KCl, K₂SO₄, etc.) and subsequent analysis of
102 its elemental composition using combustion or oxidization. However, the molecular
103 composition cannot be detected using this method (Jones and Willet 2006). Untargeted
104 metabolomics is rapidly gaining attention, but its results are highly dependent on the
105 extraction method and detection instrument used. Gas chromatography/mass
106 spectrometry (GC/MS) and liquid chromatography/mass spectrometry (LC/MS) are the
107 most accessible and widely-used methods due to their broad analytical scope (alcohols,
108 fatty acids, sterols, carbohydrates, amino acids, etc), low operation cost, and availability
109 of spectral databases of various metabolites (Brailsford et al. 2019; Brown et al. 2021;
110 Liu et al. 2021; Swenson et al. 2015). Other available methods include capillary
111 electrophoresis/mass spectrometry (CE/MS) (Warren, 2020) and Fourier transform ion
112 cyclotron resonance/mass spectrometry (FTICR/MS) (Hirai et al. 2004), which are not
113 extensively used. The compounds detected vary with the detection method used, and
114 the most suitable method that reflects soil microbe-accessible metabolites is still
115 unknown.

116 Microorganisms are the most sensitive soil quality indicators and respond quickly
117 to changes in soil DOM under chemical and organic fertilizer application (Ma et al.

118 2020b). A shift in microbial community composition indicates a change in the
119 metabolism and function of the community in a soil ecosystem (McGuire and Treseder,
120 2010). Moreover, the microbial community strongly drives the utilization and
121 mineralization of organic C and N (Ma et al. 2018). Nutrient (C, N and P) enrichment
122 induces significant changes in the soil metabolite profile and microbial C partitioning.
123 A recent study based on UHPLC-MS/MS found that inorganic nutrient enrichment
124 causes substantial shifts in both secondary and primary metabolism and changes in
125 resource flow and soil functioning, and the microbial community composition showed
126 significant metabolic flexibility (Brown et al. 2022). The systematic coupling of the
127 microbial community and soil metabolomics can provide valuable information to
128 improve our understanding of microbial strategies in response to environmental stress
129 (Swenson et al. 2018). However, the elucidation of this link is difficult due to a large
130 number of metabolites and the complexity of the microbial community.

131 Therefore, in this field-based study, we aimed to improve our understanding of
132 soil metabolic processes by exploring the response of soil metabolites to long-term
133 fertilizer application. We hypothesised that (1) the total DOM detected by traditional
134 methods, GC-TOFMS/MS and UHPLC-QTOFMS/MS, should be positive related to
135 each other; (2) soil metabolomics and the microbial community are systematically
136 coupled.

137

138 **Materials and methods**

139 *Experimental site and treatments*

140 Soil samples were collected in June 2018 from the long-term Woburn Organic
141 Manuring experiment running since 1964 in Southeastern England
142 (www.era.Rothamsted.ac.uk/WoburnFarm) to test the effects of organic manures and
143 chemical fertilizers on soil fertility and crop production. The soil is derived from Lower
144 Greensand parent material and is classified as a sandy loam-textured brown sand (10%
145 clay, 6% silt, and 80% sand, excluding organic matter content). The soil samples were
146 collected from three typical treatments that reflected current agronomic regimes: FYM
147 applied at 25–50 t ha⁻¹ y⁻¹ for 28 y (high manure application, High-M), FYM applied at
148 10 t ha⁻¹ y⁻¹ for 16 y supplemented with chemical fertilizers (low manure application,
149 Low-M), and chemical fertilizers only (No-M), with P and K inputs equivalent to 25–
150 50 t ha⁻¹ y⁻¹ FYM. Each treatment consisted of four replications. Each plot was 8.83 ×
151 8.00 m with a 5-year arable rotation (since 2003 this has been spring barley and mustard,
152 winter beans, winter wheat, forage maize and mustard, winter rye) since 1964.

153 The treatment plots received chemical fertilizers or organic manures for three
154 periods between 1964 and 2018. In the High-M treatment, FYM was applied from
155 1964–72, 1981–87, and 2003–18 (28 y in total). FYM was applied at 50 t ha⁻¹ in the
156 first two build-up periods and 25 t ha⁻¹ in the final period. In the Low-M treatment,
157 FYM was applied at 10 t ha⁻¹ from 2003 onward (16 y in total). Before this, it received
158 chemical fertilizers (P & K) equivalent to 7.5 t ha⁻¹ straw input, containing
159 approximately 30.8 kg N ha⁻¹. The No-M treatment received chemical fertilizers as N,
160 P, and K at rates equivalent to High-M during the same years. Since 2003, all treatments

161 received annual N (nitrochalk), P (triple superphosphate), and K and S (potassium
162 sulfate) fertilizers at 165 (equivalent annual rate for a 5-year crop rotation), 20, 83, and
163 36 kg ha⁻¹, respectively. All other aspects of agronomic management, including
164 harvesting, tillage regime, herbicides (including spring-applied Atlantis at 400 mL ha⁻¹,
165 Hiatus at 50 g ha⁻¹, and Sprinter at 2 L ha⁻¹) and fungicides (including spring-applied
166 Keystone at 500 mL ha⁻¹, Folicur at 800 mL ha⁻¹, and Cello at 630 mL ha⁻¹), were
167 consistent among the three treatments. The total N, P, and S inputs during the build-up
168 phase (1964–2018) under No-M were 2.46, 1.77, and 0.96 t, respectively. The total C,
169 N, P, and S inputs under High-M were 112.50, 5.80, 1.26, and 1.22 t, respectively, while
170 that under the Low-M treatment was 14.10, 2.63, 1.69, and 1.00 t. Further details of the
171 agronomic regime and experiment can be found in Ma et al. (2020b).

172 From each of four plots per treatment, the topsoil (0–23 cm plow layer) and subsoil
173 (23–38 cm) samples were collected using a 2.5 cm diameter corer (18 cores per plot to
174 make up one replicate). Winter rye (*Secale cereale* L.) was sown in the plots in 2018,
175 and sampling was performed at the grain-filling stage. The soil was thoroughly mixed
176 by hand and passed through a 5 mm sieve to any remove roots, stones, and earthworms.
177 The soil samples were then portioned into three parts: the first was stored at –80 °C to
178 analyse soil metabolites and microbial community; the second was stored at 4 °C to
179 assess soil microbial biomass; and the third was air-dried to determine basic soil
180 properties.

181 *Determination of soil properties*

182 Basic soil properties were determined using traditional methods. Soil pH was
183 determined at a 1:2.5 (v/v) soil: H₂O ratio. Total C and N were measured by dry
184 combustion of finely milled soil using a CHN-2000 Analyser (Leco Co., St. Joseph,
185 MI, USA). To determine the K₂SO₄ extractable C and N (total, organic, NO₃⁻, and
186 NH₄⁺), 5 g of moist soil was extracted with 25 mL of 0.5 M K₂SO₄ for 30 min at 200
187 rpm, and centrifuged for 10 min at 12 000 × g at 25 °C. The dissolved organic C (DOC)
188 and total dissolved N (TDN) in the extracts were detected using a multi N/C 2100S
189 TOC-TN Analyser (Analytic Jena AG, Jena, Germany). The NO₃⁻ and NH₄⁺ content in
190 the extracts were detected colorimetrically using a microplate spectrophotometer
191 (BioTek Instruments Inc., Winooski, VT, USA). Extractable organic N was calculated
192 by subtracting the NO₃⁻ and NH₄⁺ content from TDN. Soil microbial biomass C (MB-
193 C) and N (MB-N) were determined using the CHCl₃ fumigation-extraction method
194 (Vance et al. 1987). Organic C and N were extracted and detected from the fumigated
195 soil in the same manner as from non-fumigated soil. MB-C and MB-N were calculated
196 by a conversion factor of 2.22 for both C and N (Vong et al. 2003). Total soluble protein
197 in the extracts was estimated by the acid hydrolysis of proteins in solution with the
198 subsequent determination of the amino acids as described by Roberts and David (2008)
199 and have been reported previously (Ma et al. 2020b). To quantify the fraction of
200 peptides and free amino acids, the 0.5 M K₂SO₄ extracts were passed through a 1 000
201 MW ultrafiltration membrane using an Amicon[®] stirred cell (Merck-Millipore,
202 Billerica, MA, USA). Amino acids in the flow-through were detected using the
203 fluorometric OPAME method before and after acid hydrolysis with 6 M HCl (105 °C,

204 16 h) under N₂ (Jones et al. 2002).

205 *Untargeted metabolomics detected by GC-TOFMS/MS*

206 The soil samples stored at –80 °C were freeze-dried using an Edwards Super Modulyo
207 freeze-drier (SciQuip Ltd., Shropshire, UK) for 3 d. The dried soil was ground using a
208 ball mill (Retsch MM200, GmbH, Haan, Germany) to promote metabolite recovery
209 from the microbial biomass (Wang et al. 2015). The samples were extracted by 3:3:2
210 (v/v/v) acetonitrile-isopropanol-water (Brailsford et al. 2019; Brown et al. 2021), as
211 this extraction method can extract a broad range of metabolites. The untargeted
212 metabolome was analysed at the UC Davis West Coast Metabolomics Facility using an
213 automated linear exchange-cold injection system (ALEX-CIS) GC time of flight (TOF)
214 MS (Brailsford et al. 2019; Brown et al. 2021). Briefly, 0.5 µL of the extracted solution
215 was injected into an Rtx-5Sil MS capillary column (0.25 µm 95% dimethylsiloxane/5%
216 diphenylpolysiloxane coating; 30 m length × 0.25 mm i.d.; Restek Corp., Bellefonte,
217 PA, USA). This chromatography method yields excellent retention and separation of
218 primary metabolite classes (amino acids, hydroxyl acids, carbohydrates, sugar acids,
219 sterols, aromatics, nucleosides, amines, and miscellaneous compounds) with narrow
220 peak widths of 2–3 s and very good within-series retention time reproducibility of better
221 than 0.2 s absolute deviation of retention times. The GC thermal program was running
222 at 50 °C for 1 min, then increased to 330 °C at 20 °C min⁻¹, and finally maintained at
223 330 °C for 5 min, with a He mobile phase. Upon elution, samples were injected into a
224 Pegasus IV GC-TOF-MS (Leco Corp., St Joseph, MI, USA), using a mass resolution

225 of 17 spectra s^{-1} , from 80–500 Da, at -70 eV ionization energy and 1800 V detector
226 voltage, with a 230 °C transfer line and 250 °C ion source (Withers et al. 2020). A
227 mixture of internal retention index markers was prepared using fatty acidmethyl esters
228 of C8, C9, C10, C12, C14, C16, C18, C20, C22, C24, C26, C28, and C30 linear chain
229 length, dissolved in chloroform at concentrations of 0.8 mg ml^{-1} (C8–C16) or 0.4 mg
230 ml^{-1} (C18–C30) as detailed shown in Fiehn et al. (2008). The Raw data files are
231 preprocessed directly after data acquisition and stored as ChromaTOF-specific *.peg
232 files. ChromaTOF vs. 2.32 (Leco Corp.) is used for data preprocessing without
233 smoothing, with a 3 s peak width, baseline subtraction just above the noise level, and
234 automatic mass spectral deconvolution and peak detection at signal/noise levels of 5:1
235 throughout the chromatogram. Apex masses are reported for use in the BinBase
236 algorithm. The results are exported to a data server with absolute spectra intensities and
237 further processed by a filtering algorithm implemented in the metabolomics BinBase
238 database as shown in Withers et al. (2020). Both known and unknown compounds were
239 analysed using MetaboAnalyst v4.0 (Chong et al., 2018; Xia and Wishart, 2016). Prior
240 to analysis, the data were \log_{10} transformed and scaled by Pareto scaling (Chong et al.
241 2018).

242 *Untargeted metabolomics detected by UHPLC-QTOFMS/MS*

243 Complex lipid extraction was conducted using a modified bi-phasic method (Matyash
244 et al. 2008), which is advantageous as the lipids are retained in the upper extraction
245 phase, and the Methyl tertiary-butyl ether (MTBE) solvent has a density lower than that

246 of water. Compared to chloroform (CHCl₃), MTBE can be detected directly without the
247 risk of contamination from the interphase or aqueous phase. Briefly, 225 μL of MeOH
248 with internal standards was added to a 40 mg freeze-dried and ground soil sample and
249 vortexed for 20 s; 750 μL MTBE was subsequently added and vortexed for 10 min.
250 Samples were placed in a bead grinder for 30 s and then shaken for 6 min at 4 °C; 188
251 μL of MS-grade water was added, and the sample was centrifuged for 2 min at 14 000
252 × g at 4 °C. The upper phase was transferred to two tubes (350 μL/tube), and one tube
253 was evaporated to dryness using a SpeedVac. Dried extracts were re-suspended with a
254 mixture of 1:9 toluene: MeOH (v/v) and an internal standard. The samples were
255 analysed using an Agilent 1290 Infinity liquid chromatography (LC) system (G4220A
256 binary pump, G4226A autosampler, and G1316C Column Thermostat) coupled to an
257 Agilent 6530 MS (positive ion mode). Lipids were separated on an Acquity ultra high-
258 pressure chromatography (UHPLC) CSH C18 column (1.7 μm; 100 × 2.1 mm) (Brown
259 et al. 2021). The data were processed by the mass spectrometry-data independent
260 analysis (MS-DIAL) software (Tsugawa et al. 2015), followed by data clean-up using
261 the mass spectral feature list optimizer (MS-FLO) (Defelice et al. 2017). Peaks were
262 annotated and the MassHunter Quant software was applied to verify peak candidates
263 (Brown et al. 2021). To increase overall peak annotations, valid and reproducible peaks
264 were analysed using targeted MS/MS. In addition, 9 internal standards were used to
265 convert peak heights into good estimates of absolute (micromolar) concentrations for a
266 range of biogenic amines typically detected in biofluids and tissues (shown in
267 supplement materials). Notably, internal standards were included, but only for peak

268 correction and quality control. Therefore, the data presented are qualitative, and the
269 compounds were tentatively identified in line with typical untargeted analyses (Brown
270 et al. 2021). This UHPLC-TOFMS/MS method yields an excellent retention and
271 separation of acylcarnitines, trimethylamine oxide, cholines, betaines, S-adenosine
272 methionine, S-adenosine-L-homocysteine, nucleotides and nucleosides, methylated
273 and acetylated amines, di- and oligopeptides, while also yielding an excellent retention
274 and separation of metabolite classes (biogenic amines, cationic compounds) with
275 narrow peak widths of 5–20s.

276 *Soil DNA extraction and sequencing of bacteria and fungi*

277 DNA from soil subsamples (0.5 g) was extracted using a FastDNA SPIN kit (MP
278 Biomedicals, Irvine, CA, USA) following the manufacturer's protocols. A NanoDrop
279 ND-1000 UV-Vis spectrophotometer (NanoDrop Technologies, Wilmington, DE, USA)
280 was then used to identify the concentrations and quality of the extracted DNA. Primers
281 515F-806R (Brown et al. 2021) for bacteria and ITS1F-ITS2 (Gardes and Bruns 2010)
282 for fungi were used for amplification. The polymerase chain reaction products were
283 sequenced using the Illumina Novaseq platform. Bacterial and fungal sequence data
284 were processed using an in-house pipeline (Kai et al. 2017). Sequences with a length
285 exceeding 200 bp were retained for downstream analyses. Operational taxonomic units
286 (OTUs) were clustered at a 97% similarity. We annotated the taxonomic data for
287 representative sequences of bacteria and fungi using the SILVA (Quast et al. 2012) and
288 UNITE (Nilsson et al. 2019) databases, respectively. A total of 1 790 490 and 1 616

289 428 high-quality bacterial and fungal sequences were generated with an average read
290 count of 74 604 (from 55 781 to 85 255) and 67 351 (from 43 939 to 81 715) per sample,
291 respectively.

292 *Data and statistical analysis*

293 All statistical analyses were performed using R (version 3.4.3). The metabolomics data
294 were log₁₀ transformed. Agglomerative hierarchical clustering analyses were performed
295 for the metabolite concentration data under fertilizer treatment and soil depth according
296 to Pearson correlation coefficients. The dendrograms were combined with heat maps
297 generated based on the *z*-scores of metabolite concentrations. Principal component
298 analysis (PCA) was performed to determine the relationship between fertilizer
299 treatment and C, N, and metabolites at two soil depths. One-way ANOVA and Tukey
300 *post-hoc* testing were used to assess the differences among the fertilizer treatments, and
301 the Shapiro-Wilk test was used to check for normality; the top- and sub-soils were
302 analysed separately ($p < 0.05$). A random forest analysis was performed using the
303 ‘randomForest’ R package of the Linear discriminant analysis effect size (LEfSe) on
304 the Galaxy platform (Afgan et al. 2018). The interaction between metabolite
305 concentrations and the microbial community was visualised using the ‘psych’ package
306 in R and Gephi (<http://gephi.github.io/>).

307

308 **Results**

309 *Effect of long-term fertilization on soil properties*

310 In the collected sandy soil samples, manure application increased the total and dissolved
311 contents of C (Total C, DOC) and N, which increased with the FYM application rate
312 (Fig. S1). Generally, the total and dissolved C and N contents were greater in the topsoil
313 than in the subsoil. The peptide and amino acid contents were clustered with DOC. In
314 addition, MB-C and MB-N were clustered with total C and N, SOM, and protein content.

315 *Effects of long-term fertilization on primary metabolites detected by GC-TOFMS/MS*

316 The untargeted primary metabolomics analysis using GC-TOFMS/MS tentatively
317 identified 186 compounds, of which 71 were previously identified. Among the known
318 compounds, the concentrations of 33 compounds differed significantly between
319 treatments ($p < 0.05$) (supplement materials). In contrast to what was observed for the
320 total and DOC content, the dissolved SOM content extracted by 3:3:2 (v/v/v)
321 acetonitrile-isopropanol-water was generally smaller in the subsoil than that in the
322 topsoil. There were two distinct responses: the concentrations in the first group
323 decreased with long-term Low-M and High-M treatments and showed greater
324 concentrations in the topsoil compared to those in the subsoil ($n = 12$); the second group
325 had greater concentrations in the subsoil than those in the topsoil ($n = 59$). The top 50
326 most significant known metabolites identified by ANOVA are presented in Fig. 1.

327 *Effects of long-term fertilization on primary metabolites detected by UHPLC-* 328 *QTOFMS/MS*

329 The curated complex lipid analysis identified 2 944 individual compounds, of which
330 144 were known (supplement materials). Among these previously identified

331 compounds, 90 were selected due to their high concentrations. The selected compounds
332 were clustered into three groups: (1) compounds that appeared at greater concentrations
333 in the topsoil than that in the subsoil, showing greater concentrations under the No-M
334 than the Low-M and High-M treatments (n = 35); (2) compounds that appeared at
335 greater concentrations in the topsoil than that in the subsoil, showing the greatest
336 concentrations under the greater manure application (n = 24); and (3) compounds with
337 greater concentrations in the subsoil than that in the topsoil (n = 31). The top 50 most
338 significant known metabolites identified by ANOVA are presented in Fig. 2.

339 *PCA analysis of soil properties and soil metabolomics*

340 We observed a significant difference between the properties of topsoil and subsoil of
341 the Low-M treatment, and large differences between No-M and High-M treatments.
342 PCA indicated that the No-M and High-M treatments significantly influenced the soil
343 metabolomes detected by GC-TOFMS/MS and UHPLC-QTOFMS/MS (Fig. 3).

344 *Linear relationship between dissolved organic matter and metabolites detected by GC- 345 TOFMS/MS and UHPLC-QTOFMS/MS*

346 The metabolite profiles detected by GC-TOFMS/MS and UHPLC-QTOFMS/MS were
347 inversely correlated (Fig. S2). Therefore, while the metabolites detected by UHPLC-
348 QTOFMS/MS were positively correlated, those detected by GC-TOFMS/MS were
349 inversely correlated to SOM, EON, and MB-C contents (Fig. 4). In addition, several
350 compounds were detected by both GC-TOFMS/MS and UHPLC-QTOFMS/MS, such
351 as tyrosine, glucose-1-phosphate, leucine, glutamine, and isoleucine, but only

352 isoleucine detected by GC-TOFMS/MS was positively linked with that detected by
353 UHPLC-QTOFMS/MS.

354 *Response of bacterial and fungal communities to long-term organic and inorganic*
355 *fertilizer application*

356 LEfSe analysis was performed to identify the microbial taxa that differed significantly
357 between fertilization regimes (Fig. 5). The most significant enrichment indicators were
358 identified in the High-M treatment, while the least was in the Low-M treatment. Among
359 the bacteria, indicators belonged mainly to Proteobacteria, Actinobacteria, Firmicutes,
360 and Acidobacteria, the predominant bacterial phyla (Fig. 5A). Particularly, in the Low-
361 M treatment, Nitrospirae, which are involved in soil nitrification, were enriched. In the
362 High-M treatment, the identified indicators included *Bacillus* and Proteobacteria,
363 Actinobacteria, and Firmicutes. Among the fungi, the most prominent indicators were
364 Ascomycota, Mucoromycota, and Aphelidiomycota, the predominant fungal phyla (Fig.
365 5B). Long-term high-rate manure application (High-M) significantly increased the
366 abundance of Ascomycota, whereas long-term mineral fertilizer application (No-M)
367 significantly enriched Mucoromycota.

368 *Microbial community succession is driven by metabolites*

369 The random forest analysis revealed the relative importance of metabolites in
370 determining microbial community succession. The 15 most important metabolites are
371 presented in Fig. 6. The most important driver of both bacterial and fungal community
372 succession was 5'-Methylthioadenosine (MTA). After that, N-epsilon-Acetyllysine,

373 gamma-Glutamylleucine, Histidine, and 3-Indolepropionic acid correlated the most
374 with the bacterial community (Fig. 6A). Fungal community succession correlated most
375 strongly with 2'-O-Methyladenosine, 1,4-Cyclohexanedione, Isobutyryl-L-carnitine,
376 and Corticosterone after MTA (Fig. 6B). We constructed a co-occurrence network based
377 on the LEfSe and random forest analysis results to clarify further the correlation
378 between the microbial taxa and specific metabolites (Fig. 7). The 15 most important
379 metabolites for the two communities and the identified indicators were selected to
380 construct the co-occurrence network. There were more positive than negative
381 correlations between bacterial taxa and metabolites (69.5%), and fungal taxa and
382 metabolites (67.9%), in the network. Among the bacterial taxa in the network, the
383 family Planococcaceae and genus *Paenibacillus* showed the most correlations (8) with
384 metabolites (Fig. 7A and Table S1). In the case of metabolites, gamma-Glutamylleucine
385 had the most correlations (20) with bacterial taxa. The fungal network was simpler, with
386 fewer nodes and total degrees (Fig. 7B and Table S2) than the bacterial network.
387 *Aspergillus caesiellus* and *Thermomyces lanuginosus* had the most correlations (8) with
388 metabolites among the fungal taxa in the network, and MTA had the most links with
389 fungal taxa.

390

391 **Discussion**

392 *Effect of long-term organic and inorganic fertilizer application on soil organic matter*

393 Expectedly, long-term FYM increased the stock of soil total and DOM directly by

394 adding large amounts of organic C and nutrients, and indirectly by increasing the
395 microbial biomass (Liu et al. 2020; Ma et al. 2018). Organic C can be rapidly utilised
396 by microorganisms, and the microbial necromass contributes greatly to SOC (soil
397 organic C) sequestration, especially in soils supplemented with manure having
398 enhanced microbial biomass (Cui et al. 2020; Ma et al. 2020a; Wang et al. 2021).
399 Microbial necromass contributed to approximately half of the soil organic C in
400 grassland and cropland soils, based on the evaluation of glucosamine and muramic acid
401 from bacterial and fungal necromasses (Wang et al. 2021). Therefore, the increased
402 microbial biomass after FYM application can stimulate the formation of SOM. Long-
403 term high FYM application increased the EON content in the subsoil but not in the
404 topsoil, which is in direct contrast to the effect of chemical fertilizers. We ascribe this
405 to the blockage of sorption sites by organic acids and humic substances released from
406 the manure (Haynes and Mokolobate 2001), which increases soluble organic N leaching
407 to the subsoil (similar to that of soil soluble organic P) (Ma et al. 2020a). The sandy
408 soil we studied has a smaller adsorption ability compared to soils with high clay content;
409 therefore, leaching has a greater effect on dissolved SOM content.

410 *Effect of long-term organic and inorganic fertilizer application on soil metabolites*
411 *detected by GC-TOFMS/MS*

412 Besides the basic chemical and physical soil characteristics, metabolic profiles are an
413 important indicator of soil quality and ecosystem function (Withers et al. 2020).
414 Metabolites can be sensitive to changes in the soil environment condition, directly

415 related to the physicochemical properties and microbial community. The metabolomics
416 data detected by GC-TOFMS/MS was negatively linked to dissolved organic C and N
417 contents. The total metabolite content detected by GC-TOFMS/MS and UHPLC-
418 QTOFMS/MS were negatively correlated. While GC-TOFMS/MS can detect numerous
419 primary metabolites, it is generally limited by its poor resolving power for highly labile
420 metabolites and several N-containing metabolites, such as coelute and other sugar
421 compounds with the same m/z (Brown et al. 2021). Additionally, the samples were only
422 detected by MS in positive ion mode, and therefore the compounds only detectable in
423 the negative mode were missed. Furthermore, some compounds, such as glycine betaine,
424 are not amenable to derivatization, and hence are undetectable (Brown et al. 2021).
425 Therefore, in this study, the compounds detected by GC-TOFMS/MS were not
426 exhaustive, and the metabolomics data detected by GC-TOFMS/MS was negatively
427 correlated to EON. The extraction solution might also greatly affect the metabolites
428 detected. Extractions using 3:3:2 (v/v/v) acetonitrile-isopropanol-water reportedly
429 cover a broad range of metabolites, which is still lower than that when using water or
430 other solutions (Lee et al. 2012; Swenson et al. 2015). Likewise, when focusing on
431 sterols and fatty acids, greater concentrations of organic solvent are needed, and
432 aqueous solutions are better at extracting polar and small compounds due to the polar
433 nature of the compounds (Swenson et al. 2015). Our results suggest that the
434 metabolome detected by GC-TOFMS/MS might not accurately reflect the state of the
435 soil and that UHPLC-QTOFMS/MS may yield more informative results in these sandy
436 soils. This result is based on one study site, and the results may be different if focusing

437 on different soils.

438 *Effect of long-term organic and inorganic fertilizer application on soil metabolites*
439 *detected by UHPLC-QTOFMS/MS*

440 The selected compounds detected by UHPLC-QTOFMS/MS were clustered into three
441 groups. The first group included compounds that were more concentrated in the topsoil
442 than the subsoil, and more concentrated under mineral fertilizer application (No-M)
443 than under low and high manure application (n = 35). Their smaller concentration in
444 the subsoil could be due to the greater absorption by soil particles, as limited
445 compounds in the topsoil leached to the subsoil. The group was comprised of almost
446 all large molecular compounds, such as corticosterone, phenylacetamide,
447 coniferylaldehyde, quinolone, nicotine, and hexadecylamine, which might have derived
448 as secondary metabolites from soil microorganisms after they utilised the nutrients from
449 chemical fertilizers. The long-term use of chemical fertilizer might stimulate
450 microorganisms to synthesize those compounds and assimilate the inorganic nutrients
451 to adapt to the environmental changes caused by mineral fertilizer application. The
452 second group of compounds had the greatest concentration in the topsoil under high
453 manure application. This group might have been derived from farmyard manure or
454 microbial cycling. The last group had the greatest concentration in the subsoil, either
455 because they leached into the subsoil because of a smaller adsorption ability, or because
456 they were derived from special microorganisms in the subsoil adapted to oxygen-
457 deficient conditions (Ma et al. 2020a).

458 The metabolome detected by UHPLC-QTOFMS/MS was strongly correlated to
459 total and dissolved SOM, indicating that UHPLC-QTOFMS/MS better reflected SOM
460 content and composition, at least in this sandy bulk soil. In addition, the compounds
461 were not firmly correlated to the dissolved organic C but were strongly correlated to
462 extractable organic N. This might have been caused by the decoupling of C and N in
463 some compounds.

464 *Correlations between soil metabolism and the bacterial community*

465 Dissolved organic C, especially low molecular-weight compounds, including root
466 exudates, could be utilised directly as C sources by soil microbes (Swenson et al. 2015).
467 Therefore, soil metabolomics could improve our understanding of the coupling between
468 organic/inorganic compounds and microbial communities in the soil (Johns et al. 2017).
469 In this study, the most correlated factor for both bacterial and fungal community
470 succession was MTA, followed by N-epsilon-Acetyllysine, gamma-Glutamylleucine,
471 Histidine, and 3-Indolepropionic acid for the bacterial community (Fig. 6A); and 2'-O-
472 Methyladenosine, 1,4-Cyclohexanedione, Isobutyryl-L-carnitine, and Corticosterone
473 for the fungal community (Fig. 6B). MTA is a naturally occurring sulfur-containing
474 nucleoside, indicating that S metabolism is important for the formation of microbial
475 communities. Recently, due to a considerable decrease in sulfur dioxide emissions
476 following strict air-quality regulations, the application of fertilizers with a limited S
477 content, and a reduced S return via farmyard manure, S might be a limiting element for
478 microbial growth (Piotrowska-Długosz et al., 2017). N-epsilon-Acetyllysine is a

479 derivative of the amino acid lysine. A glutamyl-L-amino acid is obtained through
480 formal condensation of the gamma-carboxy group of glutamic acid with the amino
481 group of leucine. Indole-3-propionic acid is a bacterial metabolite that exerts
482 antioxidant and neuroprotective activities. Most of these metabolites are
483 amino acid derivatives, which can be utilised by soil microorganisms, hence regulating
484 the microbial community (Ma et al. 2021b). Maltose and sucrose are low molecular
485 compounds directly utilised as energy sources by microbes in the soil (Vives-Peris et
486 al. 2020). In particular, organic acids and sugars are the main drivers of shifts in soil
487 microbial communities in the rhizosphere and are positively or negatively correlated
488 with the relative abundances of bacteria (Song et al. 2020; Swenson et al. 2015).

489 Our results showed that metabolite profiling and high-throughput sequencing could
490 be successfully integrated. We found more positive correlations between bacterial taxa
491 and metabolites (69.5%), and fungal taxa and metabolites (67.9%), than negative
492 correlations in the co-occurrence network. The family Planococcaceae and genus
493 *Paenibacillus* showed the most correlations with metabolites among the bacterial taxa
494 in the network (Fig. 7A and Table S1). *Paenibacillus* is an important bacterium in bulk
495 soil that plays an important role in N fixation, hormone production, siderophore
496 secretion, and mineral nutrient activation (Li et al. 2021; Timmusk et al. 2005). In the
497 rhizosphere, Proteobacteria were the main utilizers of plant root exudates (Haichar et
498 al. 2008). They responded positively to low molecular-weight substances (Goldfarb et
499 al. 2011). However, Bacteroidetes is not a dominant bacterial phylum in bulk soil and

500 is found in high abundance in the rhizosphere (Alekklett et al. 2015). Therefore, it was
501 not the predominant bacterial phylum in the tested bulk soil. In the case of metabolites,
502 gamma-Glutamylleucine had the most links (20) with bacterial taxa.

503 Unlike the bacterial network, the fungal network was simpler, with lower nodes
504 and lower total degrees (Fig. 7B and Table S2). Previous studies have demonstrated that
505 fungi tend to decompose recalcitrant SOC, such as lignin and cellulose, and bacteria
506 then utilize the fungal-derived products (de Boer et al. 2005). Among the fungal taxa in
507 the network, the species *Aspergillus caesiellus* and *Thermomyces lanuginosus* had the
508 most correlations (8) with metabolites. In addition, MTA was found with the most
509 degrees with the fungal taxa. Soil microbial community structures can be achieved by
510 high-throughput sequencing. However, the actual microbial functions, such as their
511 metabolism, are difficult to obtain with soil metagenome or amplicon sequencing
512 (Jansson and Hofmockel, 2018).

513 The soil metabolome was mainly formed of organic acids, sugars, and sugar
514 derivatives, which were widely negatively correlated with bacterial alpha-diversity.
515 Compared to sugars, organic acids accounted more for bacterial community
516 compositions at high taxonomic ranks, while this was reversed at the species and genus
517 levels. Keystone species in co-occurrence network, such as *Microvirga*, *Bryobacter*,
518 and *Bradyrhizobium* were significantly correlated with organic acids and sugars (Liu et
519 al. 2020). We anticipate that these substrate-genome linkages could be further evaluated
520 and refined using other approaches. Stable isotope probing coupled with labelled DNA

521 sequencing (Orsi et al. 2016; Pepe-Ranney et al. 2016) and integrated NanoSIMS and
522 FISH imaging (Woebken et al. 2015; Fike et al. 2008) may be used to examine the
523 spatial localization of microbes and their activities (Swenson et al. 2018).
524 Complementary analyses of metabolic flux through real-time MS or NMR combined
525 with stable isotopes may also offer a deeper understanding of metabolic network
526 dynamics (Ina and David 2016; Jeong et al. 2017). A metabolomic profile alone cannot
527 provide a complete understanding of interacting molecular pathways and their modes
528 of regulation; the variation of metabolite levels cannot definitively infer functional
529 change. The combination of genomic and proteomic or transcriptomic results with
530 metabolites may contribute toward a more holistic understanding of soil microbial
531 function and regulation (Trauger et al. 2008).

532 **Conclusions**

533 We found that long-term farmyard manure application enhanced the total and dissolved
534 soil contents of C and N. The metabolome detected by UHPLC-QTOFMS/MS was
535 linear and positively correlated to SOM, EON, and MB-C, indicating that the
536 metabolites detected by UHPLC-QTOFMS/MS reflect the soil organic matter content
537 and composition. There were more positive correlations between bacterial and fungal
538 taxa and metabolites than negative correlations in the network. The family
539 Planococcaceae and genus *Paenibacillus* showed the most correlations with
540 metabolites among the bacterial taxa in the network. The combination of genomic and
541 proteomic or transcriptomic results with metabolites may contribute toward a more

542 holistic understanding of soil microbial function and regulation. It is impossible to
543 extract all metabolites from soil and the detected metabolites depend on the extracting
544 solution; therefore, a more detailed exploration of both extraction and detection
545 methods which more accurately reflect the composition of soil compounds and its
546 turnover is needed.

547 **Acknowledgments**

548 This work was supported by the National Natural Science Foundation of China
549 (32102488, 32172674), the UK–China Virtual Joint Centre for Agricultural Nitrogen
550 [grant number CINAg, BB/N013468/1], which is jointly supported by the Newton Fund,
551 via UK BBSRC and NERC, and the Chinese Ministry of Science and Technology. The
552 field experiment is maintained as part of the Rothamsted Long-Term Experiments
553 National Capability, funded by BBSRC (BBS/E/C/000J0300). We thank the curators
554 of the Electronic Rothamsted Archive (e-RA) for access to data from the Rothamsted
555 Long-Term Experiments.

556 **Conflict of interest**

557 The authors declare no conflict of interest.

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765 **Figure captions**

766 **Fig. 1.** Heatmap of the 50 most significant known metabolites (detected by GC-
767 TOFMS/MS) identified by ANOVA. Metabolites were clustered by Pearson correlation.
768 The color of squares linking metabolites to samples ranges from blue to red, indicating
769 the number of standard deviations from the mean. No-M, mineral fertilizers without
770 application of manure; Low-M, medium application rate of manure with mineral
771 fertilizers; High-M, high application rate of manure with mineral fertilizers. 4-amino
772 acid: 4-amino butyric acid; 4-hydro acid: 4-hydroxybenzoic acid; N-acety.: N-
773 acetylmannosamine; UDP-N-acety.: UDP-N-acetylglucosamine; gly. alf. phos.:
774 glycerol-alpha-phosphate; glucose-1-phos: glucose-1-phosphate; beta-mann.: beta-
775 mannosylglycerate.

776 **Fig. 2.** Heatmap of the 50 most significant known metabolites (detected by UHPLC-
777 QTOFMS/MS) identified by ANOVA. Metabolites were clustered by Pearson

778 correlation. The color of squares linking metabolites to samples ranges from blue to red,
779 indicating the number of standard deviations from the mean. No-M, mineral fertilizers
780 without application of manure; Low-M, medium application rate of manure with
781 mineral fertilizers; High-M, high application rate of manure with mineral fertilizers.
782 Butyl: Butylisopropylamine; N-N,N-dipro: N-(4-Piperidinyl)-N,N-dipropylamine; 4-
783 Hydro: 4-Hydroxy-1-(2-hydroxyethyl)-2,2,6,6-tetramethylpiperidine; 3-indol. Acid: 3-
784 Indoleacetic acid; Arach. Dopam.: Arachidonyl dopamine; N,N-Diethyl: N,N-Diethyl-
785 2-aminoethanol; Indole-3-carbox.: Indole-3-carboxaldehyde; Guanid. acid: 4-
786 Guanidinobutyric acid; Isobu. Carni.: Isobutyryl-L-carnitine; 1,1-Dimet.: 1,1-
787 Dimethyl-4-phenylpiperazinium; 4-amino. Acid: 4-Aminobenzoic acid; 5-Methy.: 5'-
788 Methylthioadenosine; Glycer.: Glycerophosphocholine; Atrazine-desis.: Atrazine-
789 desisopropyl-2-hydroxy; 8-Oxo-2-deoxy.: 8-Oxo-2-deoxyadenosine; N-epsilon-Acety.:
790 N-epsilon-Acetyllysine; gamma-Gluta.: gamma-Glutamylleucine.

791 **Fig. 3.** Principal component analysis (PCA) of soil carbon and nitrogen content
792 detected by traditional methods (A), and metabolites detected by GC-TOFMS/MS (B)
793 and UHPLC-QTOFMS/MS (C) under long-term (1964–2018) manure and mineral
794 fertilizer applications. Prior to analysis, the data were log₁₀ transformed. No-M,
795 mineral fertilizers without manure application; Low-M, medium application rate of
796 manure with mineral fertilizers; High-M, high application rate of manure with mineral
797 fertilizers; T, topsoil; S, subsoil.

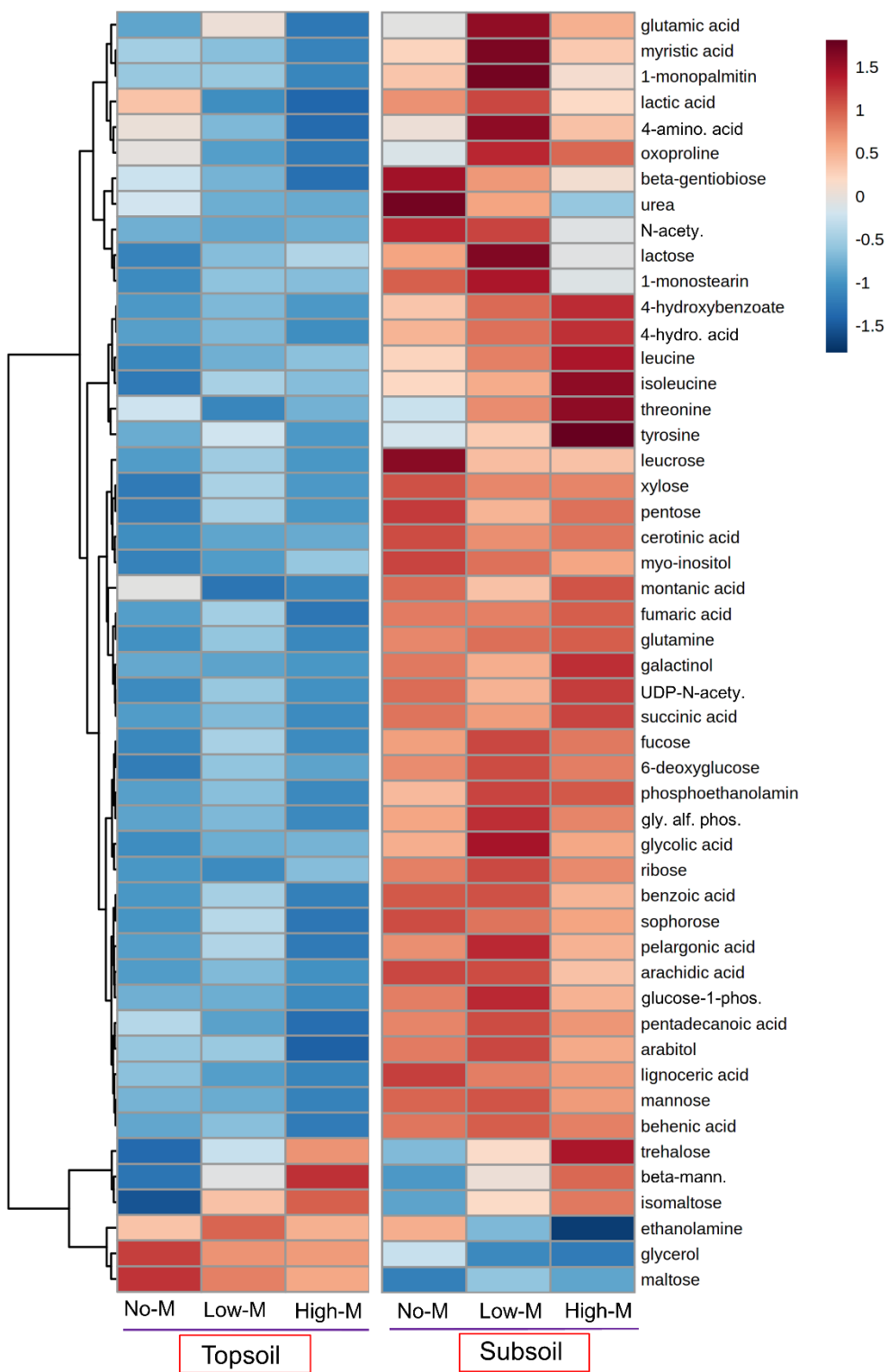
798 **Fig. 4.** Linear correlations of metabolites detected by GC-TOFMS/MS (A) and

799 UHPLC-QTOFMS/MS (B) with soil carbon and nitrogen content detected by
800 traditional methods under long-term (1964–2018) manure and mineral fertilizer
801 applications. DOC, dissolved organic carbon; SOM, soil organic matter; EON,
802 extractable organic N; MB-C, microbial biomass carbon; MB-N, microbial biomass
803 nitrogen.

804 **Fig. 5.** The response of bacterial (A) and fungal (B) communities at phylum to genus
805 levels to long-term organic and inorganic fertilizer application based on a linear
806 discriminant effect size analysis. Only taxa meeting a linear discriminant analysis
807 significance threshold of $LDA > 3$ are shown and color-coded. The six rings of the
808 cladogram indicate the domain (d), phylum (p), class (c), order (o), family (f), and genus
809 (g), from inside to outside.

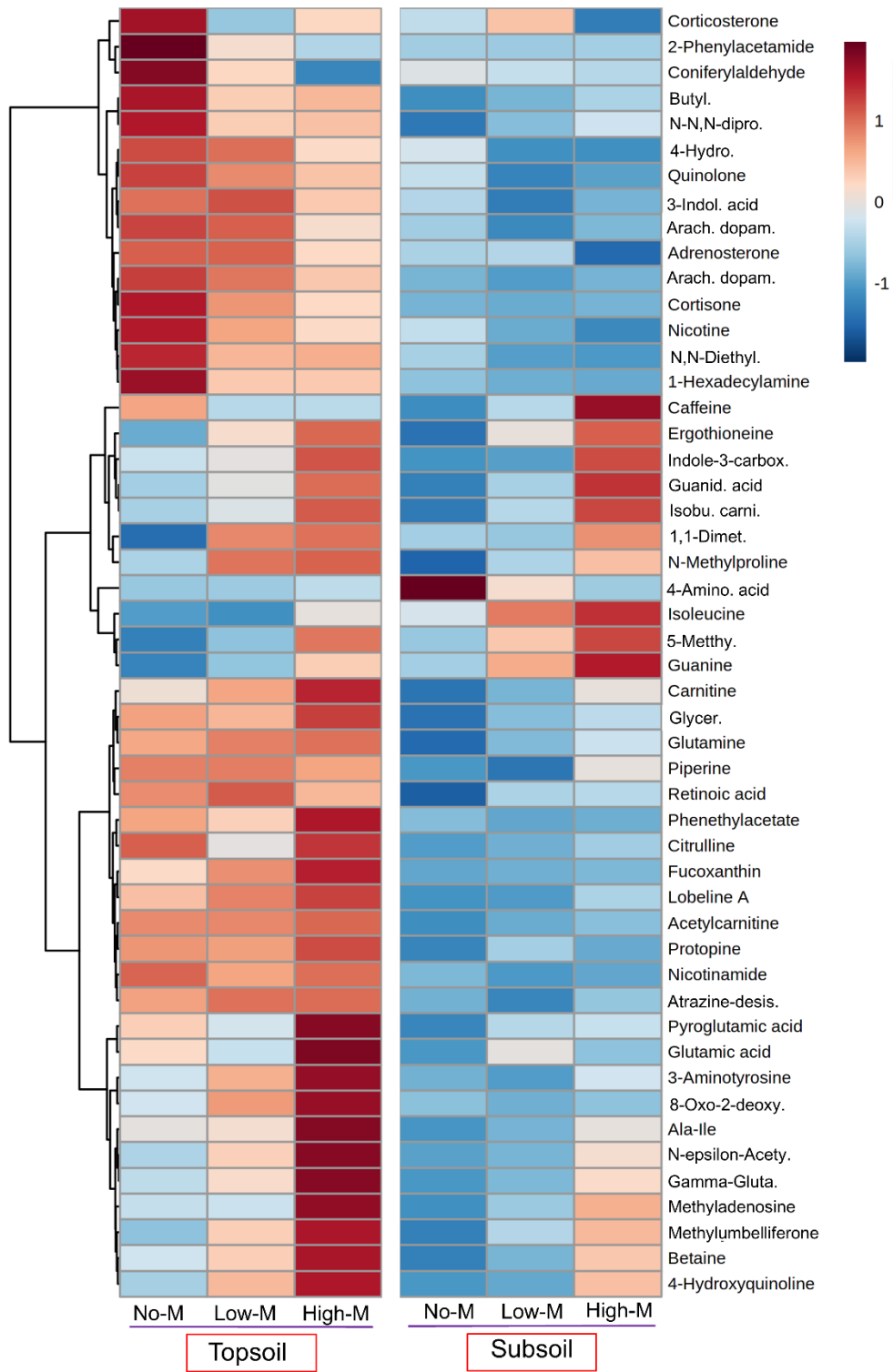
810 **Fig. 6.** Random forest analysis to determine factors affecting bacterial (A) and fungal
811 (B) community succession. The metabolites detected by UHPLC-QTOFMS/MS was
812 used in this analysis.

813 **Fig. 7.** Co-occurrence network of the metabolites and bacterial (A) and fungal taxa (B).
814 The node size represented the degree in the network. Only significant Pearson
815 correlation coefficients ($r > 0.8$ or $r < -0.8$ and $p < 0.05$) are shown. The metabolites
816 detected by UHPLC-QTOFMS/MS was used in this analysis. Light purple and red lines
817 indicate positive and negative correlations, respectively. Pink circles represent
818 microorganisms, and green circles represent metabolites.



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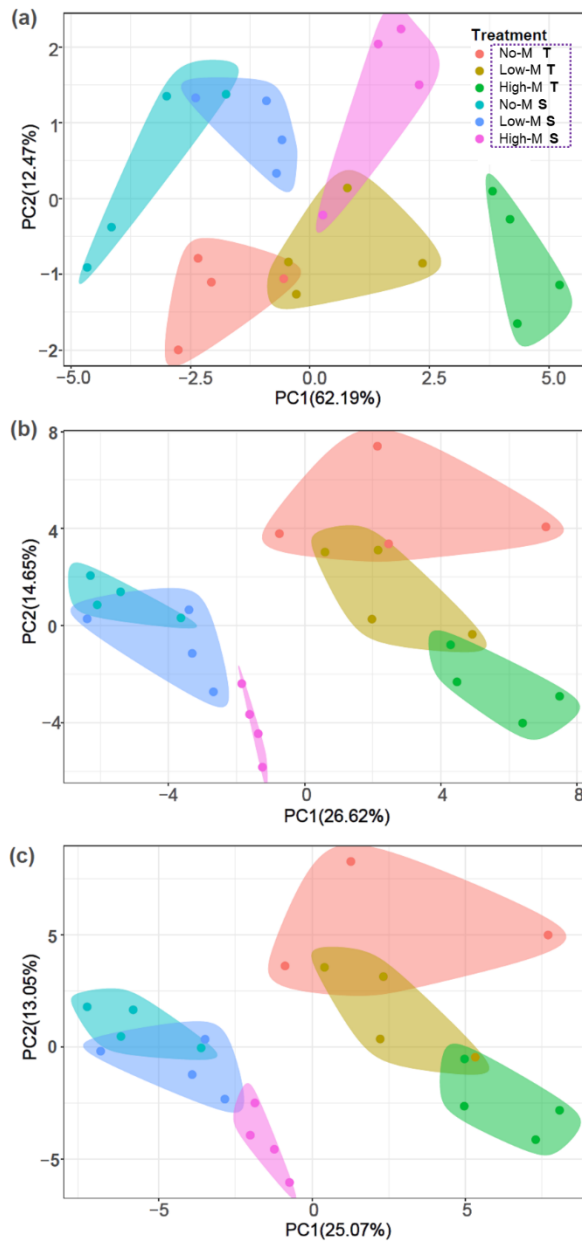
820 Figure 1



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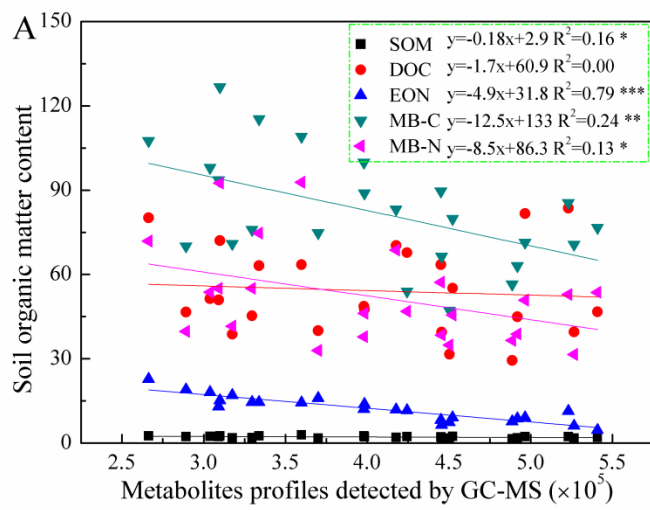
822 Figure 2

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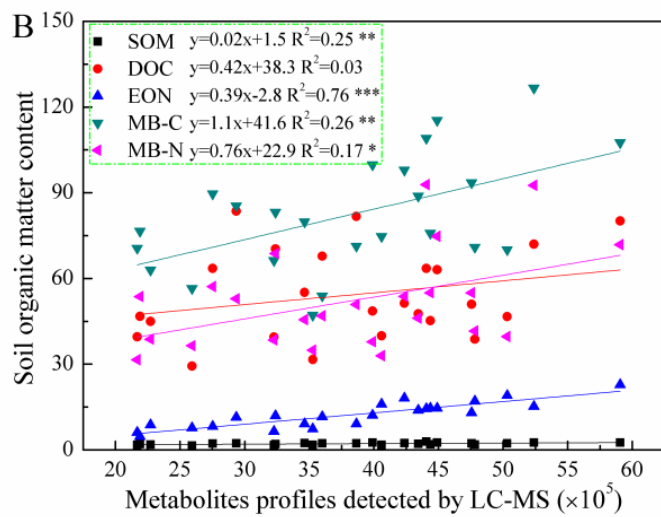


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825 Figure 3

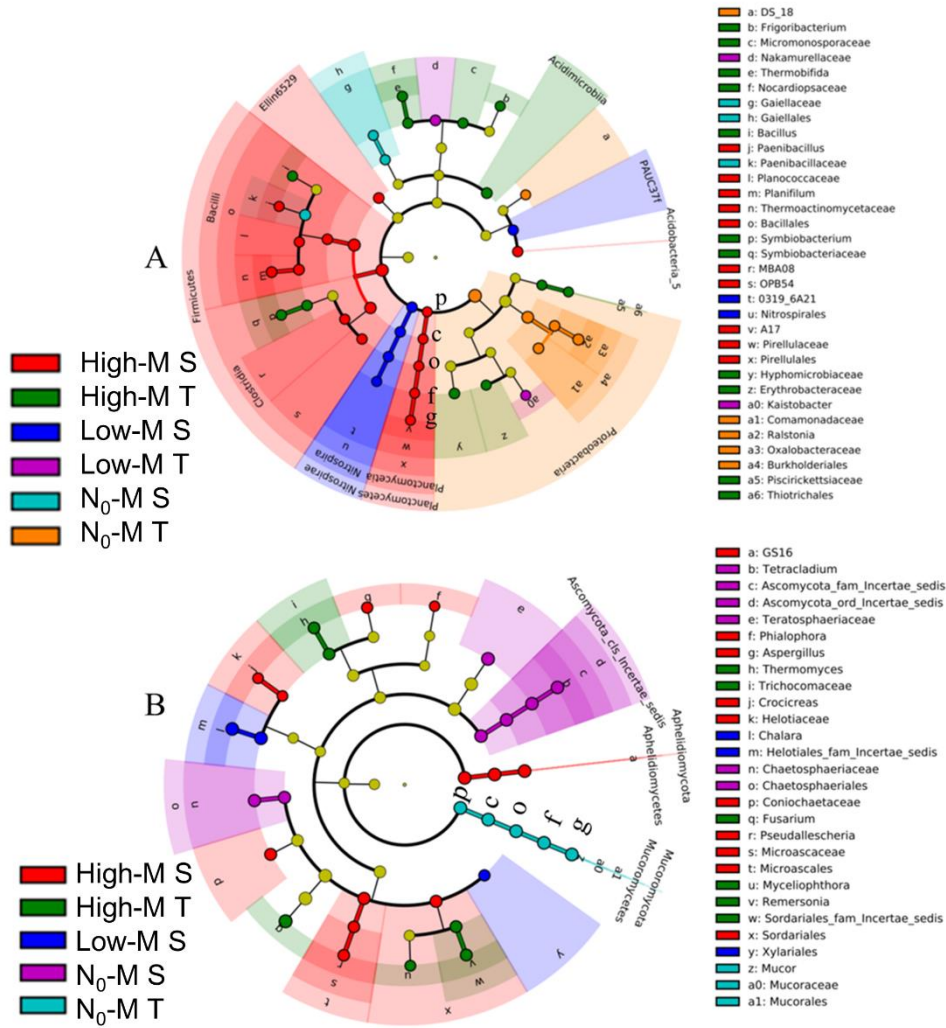


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828 Figure 4



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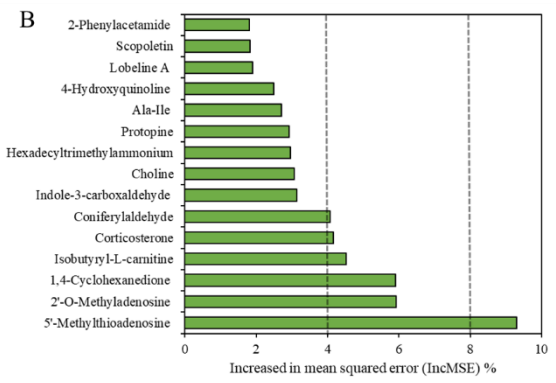
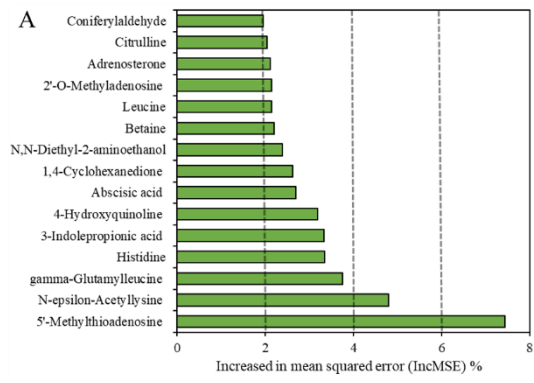
830 Figure 5

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836 Figure 6

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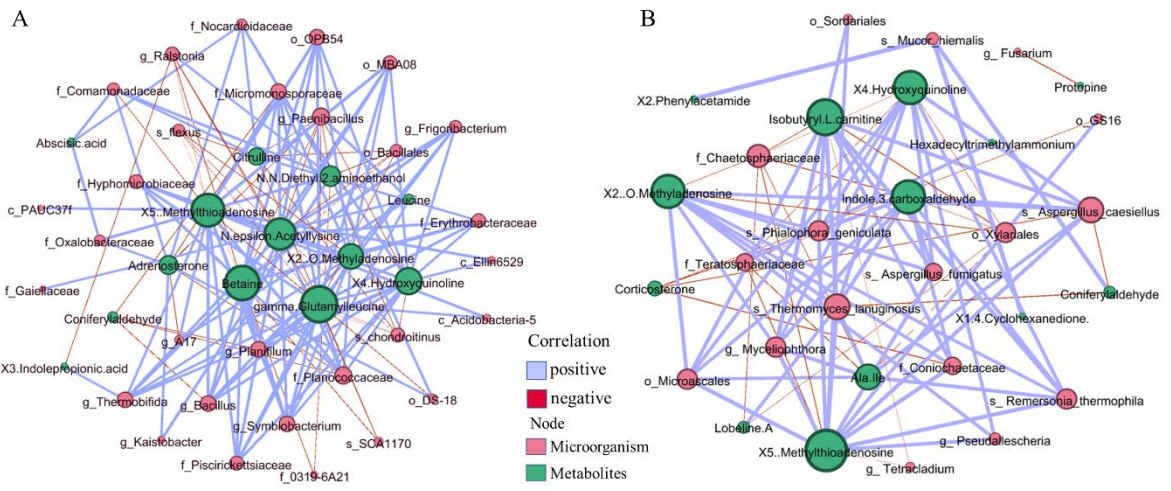
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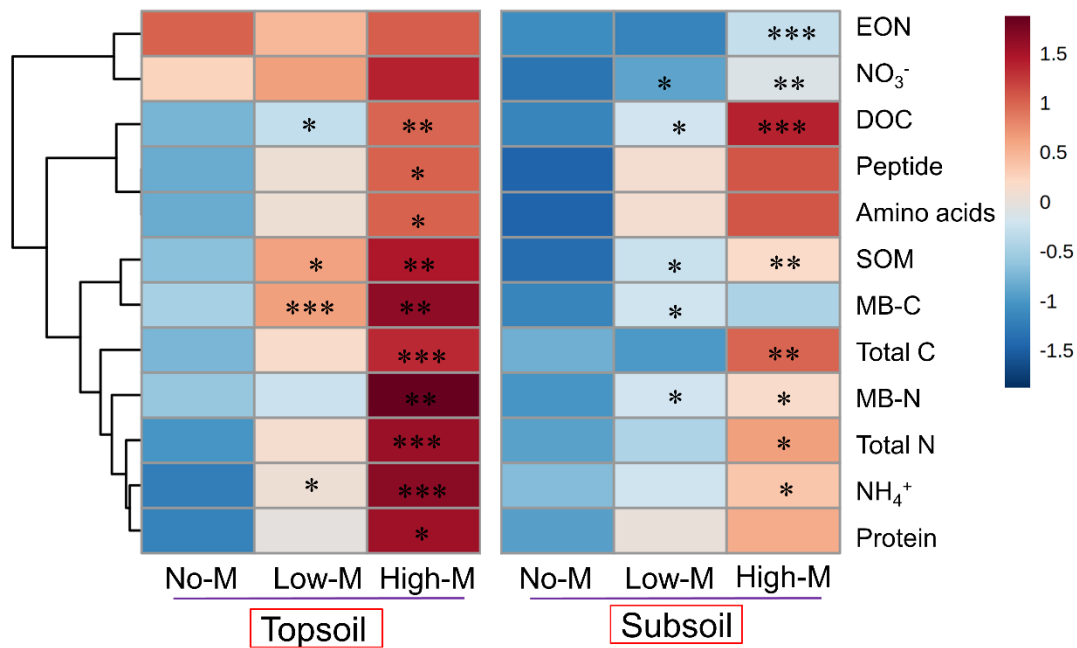
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846 Figure 7

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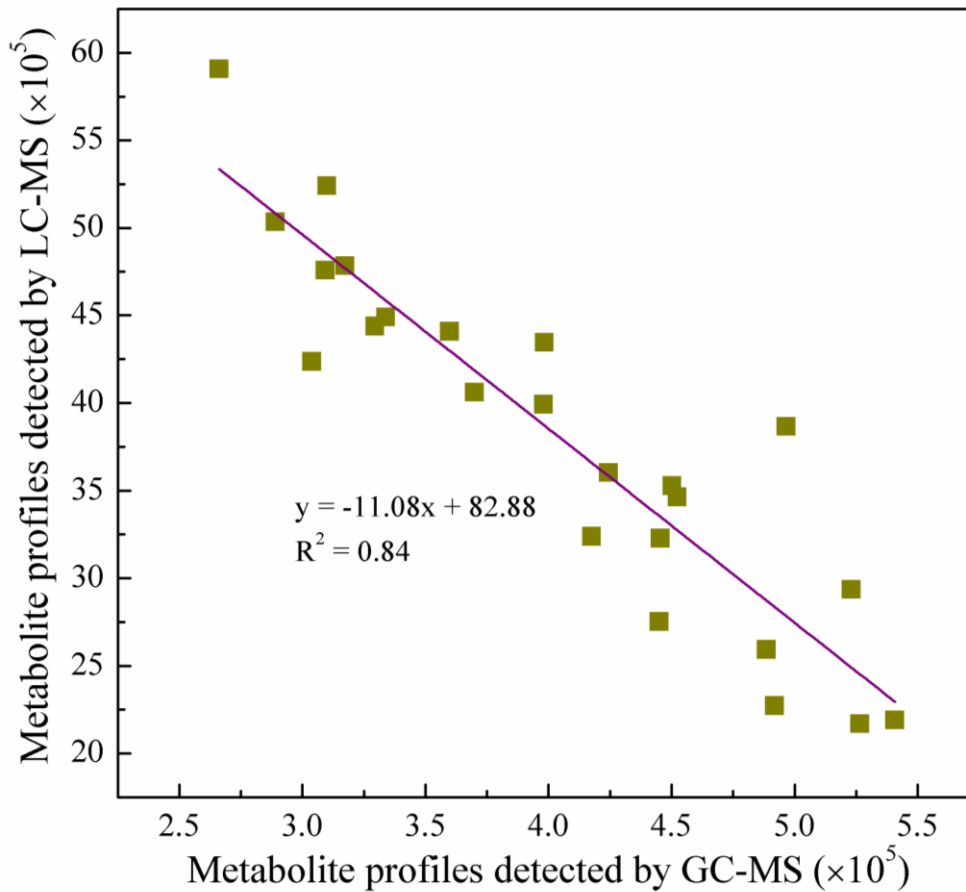
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850 Figure S1

851 **Fig. S1.** Heatmap based on hierarchical clustering of the effects of long-term (1964–
 852 2018) manure and mineral fertilizer applications on total and dissolved contents of C
 853 and N in the topsoil and subsoil. Contents that significantly decreased are displayed in
 854 blue, whereas elements that significantly increased are displayed in red. The brightness
 855 of each colour corresponds to the magnitude of the difference compared with average
 856 values. Clustering of compounds according to Pearson correlation coefficients is
 857 depicted by the dendrogram on the left. Differences between treatments (compared with
 858 No-M) in the topsoil and subsoil were separately analysed using *t*-tests. DOC, dissolved
 859 organic carbon; SOM, soil organic matter; EON, extractable organic N; MB-C,
 860 microbial biomass carbon; MB-N, microbial biomass nitrogen; No-M, mineral

861 fertilizers only; Low-M, medium application rate of manure with mineral fertilizers;
862 High-M, high application rate of manure with mineral fertilizers. *, $p < 0.05$; **, $p <$
863 0.01 ; ***, $p < 0.001$.

864



865

866 **Fig. S2.** Linear regression of metabolites detected by GC-TOFMS/MS and UHPLC-
867 QTOFMS/MS under long-term (1964–2018) manure and mineral fertilizer applications.

868

869 **Table S1** Degree of bacterial taxa and metabolites in the co-occurrence network.

Node	Label	Degree
Bacteria	f_Planococcaceae	8
Bacteria	g_Paenibacillus	8
Bacteria	f_Micromonosporaceae	7
Bacteria	g_Bacillus	7
Bacteria	g_Planifilum	7
Bacteria	g_Symbiobacterium	7
Bacteria	g_Thermobifida	7
Bacteria	o_OPB54	7
Bacteria	f_Erythrobacteraceae	6
Bacteria	f_Hyphomicrobiaceae	6
Bacteria	f_Piscirickettsiaceae	6
Bacteria	g_Ralstonia	6
Bacteria	o_MBA08	6
Bacteria	s_chondroitinus	6
Bacteria	f_Comamonadaceae	5
Bacteria	g_Frigoribacterium	5
Bacteria	o_Bacillales	5
Bacteria	s_Flexus	5
Bacteria	f_Nocardiodaceae	4
Bacteria	f_Oxalobacteraceae	4
Bacteria	g_A17	3
Bacteria	o_DS-18	3
Bacteria	s_SCA1170	3
Bacteria	c_Acidobacteria-5	2
Bacteria	c_Ellin6529	2
Bacteria	f_0319-6A21	2
Bacteria	g_Kaistobacter	2
Bacteria	c_PAUC37f	1
Bacteria	f_Gaiellaceae	1
Metabolites	gamma-Glutamylleucine	20
Metabolites	Betaine	19
Metabolites	5'-Methylthioadenosine	18
Metabolites	N-epsilon-Acetyllysine	17
Metabolites	4-Hydroxyquinoline	14
Metabolites	2'-O-Methyladenosine	13
Metabolites	N,N-Diethyl-2-aminoethanol	10
Metabolites	Adrenosterone	9
Metabolites	Citrulline	8
Metabolites	Coniferylaldehyde	4
Metabolites	Leucine	4
Metabolites	Abscisic acid	3
Metabolites	3-Indolepropionic acid	2

871 **Table S2** Degree of fungal taxa and metabolites in the co-occurrence network.

0	Label	Degree
Fungi	s_ <i>Aspergillus</i> _caesiellus	8
Fungi	s_ <i>Thermomyces</i> _lanuginosus	8
Fungi	f_ <i>Chaetosphaeriaceae</i>	7
Fungi	g_ <i>Myceliophthora</i>	6
Fungi	o_ <i>Microascales</i>	6
Fungi	o_ <i>Xylariales</i>	6
Fungi	s_ <i>Phialophora</i> _geniculata	6
Fungi	s_ <i>Remersonia</i> _thermophila	6
Fungi	f_ <i>Coniochaetaceae</i>	5
Fungi	f_ <i>Teratosphaeriaceae</i>	5
Fungi	s_ <i>Aspergillus</i> _fumigatus	5
Fungi	g_ <i>Pseudallescheria</i>	3
Fungi	s_ <i>Mucor</i> _hiemalis	3
Fungi	g_ <i>Tetracladium</i>	2
Fungi	o_ <i>GS16</i>	2
Fungi	o_ <i>Sordariales</i>	2
Fungi	g_ <i>Fusarium</i>	1
Metabolites	5'-Methylthioadenosine	14
Metabolites	Isobutyryl-L-carnitine	12
Metabolites	Indole-3-carboxaldehyde	11
Metabolites	2'-O-Methyladenosine	11
Metabolites	4-Hydroxyquinoline	11
Metabolites	Ala-Ile	8
Metabolites	Corticosterone	4
Metabolites	Coniferylaldehyde	3
Metabolites	Lobeline A	3
Metabolites	Hexadecyltrimethylammonium	1
Metabolites	Protopine	1
Metabolites	1,4-Cyclohexanedione	1
Metabolites	2-Phenylacetamide	1

872