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Impacts of changing weather patterns on the dynamics of water pollutants in agricultural catchments: Insights from 11-year high temporal resolution data analysis

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Abstract:

Widespread and long-term shifts in weather patterns are contributing to further degradation of surface water quality. This challenge caused by the increasing frequency of extreme weather events requires appropriate adaptation of current mitigation strategies. But to confirm the need to redesign such strategies, an understanding of the impacts of increasing weather extremes on pollutant losses in different catchment types is required. With this in view, this study investigated the impact of changing weather patterns on the inter-seasonal and inter-annual dynamics of nutrient losses in six agricultural catchments in Ireland over 11 years. The high temporal resolution data (10-min) from these intensively managed catchments represented different characteristics and management practices. Mann-Kendall Trend Analysis and Generalised Additive Models were used to study nutrient concentration trends, and to investigate the significance of water discharge, precipitation, potential evapotranspiration, soil moisture deficit, air temperature, and soil temperature on the losses of nutrients, respectively.

The analysis of historical data revealed changes in the trends of daily average nitrate (NO₃-N), phosphorus (P), and suspended sediment (SS) concentrations in association with significant increasing trends in air temperature, soil temperature, and precipitation across the same month over 11 years of monitoring. While discharge was significantly contributing to the concentrations of NO₃-N, P, and SS across different catchments, air and soil temperature were significantly correlated to NO₃-N losses, and precipitation was the major contributor to regulating P (total P and total reactive P) concentrations. In short, air temperature, soil

temperature, soil moisture deficit, and precipitation were the main climatic drivers regulating the nutrient concentrations while the soil chemistry and drainage status were the non-climatically related drivers.

The results revealed that the extent of the impact of climatic drivers depends on catchment characteristics. Therefore, expanding the application of this type of study would facilitate better understanding of current and future challenges to water management and provision of climate-resilient mitigation strategies for different catchment typologies.

Keywords:

Nitrate, phosphorus, suspended solids, climate change, farming, water quality

1. Introduction

Excess nutrients (i.e. nitrogen (N) and phosphorus (P)) and suspended sediments coming from agriculture can be delivered into lakes, streams, estuaries, and coastal waters causing degradation of water quality (Maúre et al., 2021; Basu et al., 2022; Beusen et al., 2022). Current European Legislation requires the application of mitigation strategies to meet the goal of the Water Framework Directive (WFD) (2000/60/EC) to achieve ‘good ecological status’ in all waterbodies by 2027. However, the overall ecological status of waterbodies has declined in some countries, including Ireland, which has reported increasing nutrient concentrations, mainly from agricultural sources, over the period 2016-2021 (EPA, 2021^a). Hence, a key challenge is the need to balance agricultural intensification for food security with achieving ‘good status’ in waterbodies (Moal et al., 2019). This is a global and growing issue which also impacts economic/social welfare as well as the sustainability of ecosystems and biodiversity (Weng et al., 2020).

To achieve an improvement in waterbody status, an understanding of both the nutrient sources (e.g., soils and farming practices), and climate drivers of water pollutant mobilisation/delivery

are required. Recent research has particularly highlighted the need for a comprehensive analysis of the links between the weather patterns associated with changing climate and river pollutant concentrations (Mellander and Jordan, 2021). Climate change is exerting extra pressures on water quality as an increasing frequency of extreme weather events (drought and flood) is being observed (Gascuel-Oudou et al., 2022). For example, reduced discharge during drought conditions can lead to the reduced dilution of pollutant discharges from point-sources and therefore can increase fluvial nutrient concentrations (Hughes et al., 2021). Changes in soil and water temperature and soil water content can also modify water quality indirectly (Seyyedhashemi et al., 2022). Drought can alter hydrological connectivity (Erhardt et al., 2021) and disconnect shallow flow pathways from the river network (Yang et al., 2018). Additionally, during the first heavy rain or so-called ‘first flush’ after drought conditions decreased denitrification and the subsequent flushing of accumulated nutrients have been observed (Outram et al., 2014; Strohmenger et al., 2020; Zhang et al., 2022; Winter et al., 2023).

According to the Irish Environmental Protection Agency (EPA, 2021^b), starting from 1980s, every decade has been warmer than the previous one and during the last ten-year decade (2010-2019), the average temperature was the warmest on record. An increase in the magnitude, frequency and duration of droughts is further expected during summer times (Merese and Murphy, 2023). While the average annual precipitation in Ireland has increased approximately 5% (60mm) during 1981-2010 compared to 1961-1900, substantial decreases in average annual spring and summer precipitation, and significant increase in frequency of extreme events in winter and autumn are expected (EPA, 2021^b). Weather extremes have been shown to have diverse impacts on nutrient loads within different agricultural landscapes depending upon their specific characteristics (Mellander et al., 2021; Ezzati et al., 2023) and boundary conditions (catchment boundaries) (Ehrhardt et al., 2021). These impacts generally correlate with changes in runoff volume (Coffey et al., 2021) and are therefore strongly controlled by precipitation and air temperature (Jordan et al., 2014; Paul et al., 2019; Ezzati et al., 2023; Hadush & Murphy, 2023). On the other hand, catchments, in a global scale, also have contrasting characteristics such as soil type and texture (including average clay content), land use (the fraction of arable land), management, and hydrology (Sandström et al., 2020) which result in different responses to the same weather events. The differences between catchments are comprised of variations in the sensitivity of hydrological responses to climatic drivers (Sulis et al., 2011), and thus, variations in intrinsic controls relevant to nutrient mobilisation and

96 delivery and differences in farming systems. Hence, nutrient dynamics in catchments, despite
97 close geographic proximity but with different controls and mitigation strategies, have shown
98 to respond differently to the same large-scale weather extremes (Mellander & Jordan, 2021).

100 Existing management efforts and mitigation strategies have not only failed to improve water
101 quality in Ireland (EPA, 2023), but are also incapable of buffering the effects of weather
102 extremes in the already-vulnerable agricultural sector at a global scale. This failure is of greater
103 importance in countries where the agri-food sector is an integral part of the economy and
104 society (Abbass et al., 2022). According to Samaniego et al. (2018), Europe may face a 40
105 (± 24) % increase in drought affected areas in the absence of effective mitigation strategies
106 during the 21st century (Caretta et al., 2022). Therefore, when considering the more frequently
107 occurring extreme weather events, more targeted approaches and resilient management
108 strategies are required to deliver catchment-specific and climate-resilient adaptation. However,
109 uncertainty remains regarding the causes and severity of water quality degradation due to
110 changing climate (Refsgaard et al., 2013).

112 The modelling of water quality, which is often used to guide policymakers in developing
113 appropriate mitigation strategies, is facing new challenges in representing nutrient trends,
114 legacies, delivery, and mobilization in view of current and projected climate change scenarios
115 (Mellander et al., 2018; Mellander et al., 2022^b). The impacts of the changing weather patterns
116 have also been less considered in policy reviews (Mellander et al., 2018). According to
117 Gascuel-Oudou et al. (2022), these challenges are due to three main reasons: 1. Lack of long
118 time series data for water quality and chemical concentrations compared to hydrologic fluxes
119 and storage measurements; 2. Lack of detailed understanding of climate-water quality
120 relationships compared to well-developed, but highly generalised, climate-hydrology models
121 at global and regional scales (Lintern et al., 2021); 3. Difficulty in distinguishing climate effects
122 from those due to changes in farming activities (e.g. types of the crops) (Lungarska and Chakir,
123 2018). In addition, seasonality drives changes in nutrient concentrations as a consequence of
124 changes in hydrologic flow pathways, climate change, and associated biological processes
125 (Covino et al., 2021). However, understanding the seasonal dynamics of nutrients is highly
126 complex and ensuring good water quality requires catchment-specific approaches (Warner et
127 al., 2021).

Changes in **the** nutrient concentrations are influenced by mobilisation and hydrological processes, driven by weather and agricultural management, and controlled by physical catchments characteristics (e.g. soil/bedrock drainage and chemistry) and can as such be subtle (Mellander et al., 2018). While nutrient concentrations can vary within any span of time, the impacts of extreme runoff events are typically more drastic and can be exhibited immediately. Changes in the nutrient loads, on the other hand, represent the cumulative quantity without capturing fluctuations, while hydrology may also override any processes in mobilisation (Mellander et al., 2022^a).

This study used 11-years (2010-2021) of sub-hourly (10-min basis) water quality (i.e., concentration data) and climatic data from six hydrologically contrasting agricultural catchments in Ireland. The objectives were to statistically assess the presence of seasonal and annual trends in nutrient and sediment concentrations using Mann-Kendall Trend Analysis, and to evaluate the impact of climatic data (i.e. precipitation, air/soil temperature, soil moisture deficit), on nutrient source loading and in view of management practices, using Generalised Additive Models. Such analysis may facilitate a better understanding to counter future challenges for water management and provision of climate- smart mitigation strategies.

2. Materials and Methods

In order to identify any existing trend in nutrient concentrations in view of recent extreme weather events in Ireland (prolonged draught periods and/or heavy precipitation), and in order to understand the importance of climatic variables in regulating nutrient losses throughout different months of the year, long-term high-resolution water quality data from agricultural catchments with contrasting physical and management factors were studied using different statistical modelling techniques.

2.1. Site study description

Six agriculturally-dominated catchments (one karst spring contribution zone and five river catchments) monitored by the Irish Agricultural Catchments Programme (Mellander et al., 2022^{a,b}) were investigated (Fig. 1 and Tab. 1). The catchments are all intensive agriculture areas and comprise a varying mix of grassland (pasture) and arable land use. They all have different physical characteristics, i.e., slopes, soil types, drainage, and areas ranging from 3 to

31 km². The base-flow (Bf_i) indices and detailed fertiliser applications rates have been assessed in Mellander et al., 2022^a.

Figure 1

Ballycanew (11.9 km²-Co. Wexford) is in the southeast of Ireland. This catchment has mostly poorly drained soil with land use dominated by grasslands (77%) grazed by dairy cows. There was a 30% increase in stocking rate (livestock per area) between 2010-2018. Grazing intensity is highest in May and fertilisers are mostly N-based. December and January are a “closed period” as defined by Nitrate Action Programme in Ireland (NAP) (DAFM, 2018), during which grazing and the spreading of fertilisers are prohibited. Excess slurry stored in tanks is typically spread on silage fields at the end of March. The hydrology is flashy (base-flow index (Bf_i) of 0.63) with dominating surface/near surface pathways (Mellander et al., 2015).

Castledockrell (11.2 km²-Co. Wexford) is also located in southeast Ireland. This catchment has mostly well-drained soils with 72% of the catchment under arable land use (66% of the total area is under tillage which is the highest of the catchments studied). Hydrological pathways are primarily subsurface (Mellander et al., 2016) with Bf_i of 0.78.

Corduff (3.3 km²-Co. Monaghan) is in northeast Ireland. This poorly drained catchment has no tillage, has low chemical fertiliser application rates (as detailed in Mellander et al., 2022^a), and is dominated by grassland for sheep and suckler cows. The catchment has a flashy hydrology (Bf_i of 0.57) dominated by surface/near surface pathways.

Cregduff (31.2 km²-Co. Mayo) is a mainly permanent grassland karst spring contribution zone, located in west Ireland. Sheep are primarily grazed and have a longer grazing period than cattle. The soil is well drained and relatively thin. Hydrological pathways are dominantly subsurface (Mellander et al., 2013) and Bf_i is 0.82. The rate of N-application in this catchment is low relative to other catchments.

Dunleer (9.5 km²-Co. Louth) is in northeast Ireland. Farm types in this catchment are mixed at approximately 40% grassland, half of which is grazed by dairy and half by beef cattle. This catchment has moderately drained soils with flashy hydrology (Bf_i of 0.66) and mostly surface/near surface pathways (Mellander et al., 2012). Chemical N fertilisers are spread in

April/March but available soil P has been notably increasing during the recent years. The ploughing/sowing take place during October and November.

Timoleague (7.6 km²-Co. Cork) is in southwest Ireland, is well-drained and dominated by grassland and subsurface hydrological pathways (Mellander et al., 2016) with B_f of 0.73. There has been an increase in livestock grazing intensity and the percentage of land under derogation (allowance to farm at livestock above the limit of 170 kg N ha⁻¹ under the EU Nitrates Directive). The P-index (soil fertility) is also increasing more than expected. Maize is one of the dominant crops which requires substantial slurry spreading typically between mid-January and October.

The management practices for a typical farming calendar in each catchment were collated from discussion with the local farm advisor and knowledge transfer expert (Fig. 2).

Figure 2

2.1.2. Hydrochemistry data collection

Bankside P analysers (Hach-Lange Sigmatex-Phosphax) (Melland et al., 2012) were located at the catchments outlets which measure total digested P (TP) and total reactive phosphorus (TRP) concentrations on unfiltered samples. The measuring range is 0.010 mg L⁻¹ to 5.000 mg L⁻¹, and the detection limit is 0.010 mg L⁻¹. Total oxidized N (TON) was monitored using Hach-Lange Nitratex SC-Plus UV instruments (co-located at the outlets). The measuring range is 0.1–50 mg L⁻¹ and assuming a low NO₂-N concentration (Melland et al., 2012), TON is considered equivalent to NO₃-N. The measuring of all nutrients was carried out based on a 10-min basis.

Stage-discharge rating curves on Corbett flat-v non-standard weirs have been established at the individual catchment outlets. The ratings are based on the velocity-area method with an OTT Acoustic Doppler Current meter (in WISKI-SKED software). An OTT Orpheus Mini vented pressure instrument installed in a stilling well adjacent to the weirs records the water level every 10 min. The river discharge is calculated by converting the water level to flow. In Cregduff, the discharge is calculated using an ultrasonic sensor (Thermo-Fisher time-of-flight area velocity) which is placed in an engineered uniform cross-section.

Suspended sediment (SS) concentrations were estimated using turbidity reading and corresponding concentration-turbidity conversion relationships (excluding the Cregduff catchment given its karst geology) (Sherriff et al., 2016).

2.1.3. Weather data collection

Precipitation, air temperature (T_{air}), soil temperature (T_{soil}), relative air humidity, solar radiation, wind speed and wind direction are measured at ten-minute intervals by a weather station (BWS200, Campbell Scientific, www.acpmet.ie) located in the central valley floor within each study catchment. The effective rainfall was calculated by subtracting potential evapotranspiration (derived from the Penman-Monteith equation (Monteith, 1965)) from measured rainfall.

The soil moisture deficit (SMD), which responds to changes in air temperature and rainfall, was calculated as the amount of rain needed to bring the soil back to field capacity using a SMD model (Schulte et al., 2005).

Table 1

2.3. Data analysis

2.3.1. Monthly trend analysis

To investigate the seasonality that may be hidden in inter-annual trends, and compare the nutrient concentrations dynamics in each month across different years, the temporal trends in average daily values of climatic-related explanatory drivers and nutrient concentrations were calculated using the non-parametric rank-based Mann-Kendall test (Kendall, 1975). The analysis was carried out over 11 years of the monitoring period. This test accounts for non-normality of hydrological (Yue et al., 2002) and climatological data (Partal and Kahya, 2006). Hence, the method provided accurate estimation and enabled comparison of any increasing or decreasing trends of variables in each particular month over more than a decade. The null hypothesis assumed that the data (x_1, \dots, x_n) consist of n independent and identically distributed random variables and $P < 0.05$ was considered as a significant trend for calculating slopes of time series (Sen, 1968). The intercept (α) was calculated as followed:

$$\alpha = x_{0.5} - s * y_{0.5} \quad \text{Eq.}$$

1

where $x_{0.5}$ and $y_{0.5}$ are median values of variables.

2.3.2. GAM analysis of significance of drivers

Generalised Additive Models (GAMs) (Hastie and Tibshirani, 1990) were fitted to the daily average time series data, in order to evaluate the significance of the impact of different climate variables on regulating water pollutant losses over different months of the year (on condition that an increasing/decreasing trend existed, see section 2.3.1). GAMs, as an extension to generalized linear models with smoothing functions, are becoming a widely applied statistical test for evaluation of changes in water quality at large scales (Murphy et al., 2019) with multiple variables (Wood, 2006) and consider uncertainty and random effects of both parametric and non-parametric variables. Hence, GAMs add flexibility to the modelling of complex non-linear relationships between response and explanatory variables. This is especially useful since water chemistry variables are influenced by each other or by the climatic drivers (von Brömssen et al., 2021, Ezzati et al., 2023).

$$F(E(y)) = f(x_1, \dots, x_n) = \beta + s_1(x_1) + s_2(x_2) + \dots + s_n(x_n) \quad \text{Eq.}$$

where y_i is the response variable for $i=1, \dots, n$ and x_1, \dots, x_n are independent variables. $E(y)$ is the expected value of y [as an alternative for the response distribution apart from the normal distribution], β is the model intercept, and $s_{i=1-n}(x)$ is a smooth function wrapping the independent variable.

The resulting accuracy in capturing the relationships between response variables (water quality dynamics, i.e. nutrient losses) and explanatory data (climatic drivers of Tair, Tsoil, Precipitation, and discharge) was checked based on the flexibility of the curve fitting to multiple variables (Ezzati et al., 2023). A low p-value indicated that residuals were not randomly distributed (hence there is enough data to capture subtle non-linear linkage between variables and showing significance of explanatory variables), and a significant downward or upward trend was identified if the entire confidence band of the derivate of the trend

smoothness was below or above zero, respectively (von Brömssen et al., 2021). Here, the “mgcv” package in the R statistical software (Wood, 2017) was applied while “gam.check” (as a tool provided by “mgcv”) was used to check model assumptions and assess effective degrees of freedom for the model terms.

3. Results

3.1. Hydrochemistry

Timoleague and Corduff had the highest annual average water discharge with 643 and 616 mm yr⁻¹, respectively. Cregduff outlet is in an estimated spring contribution zone which explains its very low discharge compared to the rest. However, among the other remaining five catchments, Dunleer had the lowest mean annual discharge (402 mm yr⁻¹) (Fig. 3). The values of the other response variables varied significantly among the six study catchments due to their diverse characteristics and management. Ballycanew had the largest mean SS concentration with a negatively skewed observed concentrations. Cregduff and Corduff had the lowest middle quartile in all water pollutant variables, while the ranges of TP and TRP in Castledockrell were as low as in those two catchments, but the NO₃-N range was the highest among all catchments. Dunleer and Timoleague had normally distributed observed concentrations in NO₃-N and TP and TRP concentrations, and a positively skewed SS with outliers.

According to the farming calendar and local catchment advisors, slurry spread during spring/summer time is more common in grassland areas where as ploughing and fertiliser application take place during spring on arable land (Fig. 2). For example, in Castledockrell, in which the NO₃-N concentration was highest among all the study catchments (Fig 3), the significant increase of this nutrient in April coincided with chemical fertiliser application and ploughing/sowing (Fig. 2).

Figure 3

3.2. Monthly trend analysis

The poorly drained catchments of Ballycanew and Corduff and the moderately-drained Dunleer had more frequently occurring monthly trend changes in the concentrations of $\text{NO}_3\text{-N}$ and P compared to the other study catchments (Fig. 4). The increase in monthly trends of nutrient losses in Ballycanew occurred during January, May, and September, while in Dunleer, the increase occurred during October-April. Corduff experienced an increase in monthly trend in nutrient losses throughout the year except in April and December (Fig. 4).

The monthly trend analysis of climatic drivers in Ballycanew and Corduff also showed more fluctuations compared to the other study catchments (Fig. 4). The air and soil temperature increased during summer in Ballycanew and Castledockrell. These two catchments are geographically close, yet, the changes in weather pattern were not identical.

There was a highly significant increase in TP and TRP concentrations leaving the Timoleague catchment during September; however, it does not coincide with a significant increase in the corresponding monthly trends of any of the climatic variables (Fig. 4).

Changes in monthly trends in climatic drivers were observed more frequently in Corduff compared to the other study catchments (Fig. 4). Dunleer experienced significant changes in potential evapotranspiration during September-March. There was also overlaps of the changes in monthly trends in nutrient concentrations with monthly trends in climatic drivers in Dunleer (during January, March, November, December).

Figure 4

3.2. Significance of drivers from GAMs

None of the climatic drivers were significant in regulating $\text{NO}_3\text{-N}$ losses during January-March, September, and December. Similarly, none of the drivers were significant in regulating TP/TRP concentration trends during April-May, and October, in any of the study catchments (Table 2 and 3). However, May-August and October-January were generally the time windows when significant changes in monthly trends occurred.

Discharge appeared to be one of the main regulators of $\text{NO}_3\text{-N}$ losses during months in which, a significant increase of $\text{NO}_3\text{-N}$ concentrations was observed. The significance of discharge occurred alongside its climatic drivers of Tair, Tsoil, and Precipitation. Tair was more

associated with increases in NO₃-N and Precipitation with P. Summer time appeared to be the period in which most of the trend changes in the concentrations of both nutrients (Figure 4) and the drivers of concentrations including climatic drivers (Table 2 and 3) took place.

In terms of NO₃-N losses, Corduff was mostly influenced by climatic drivers followed by Dunleer, Castledockrell, and Timoleague (Table 2). In terms of P (TP and/or TRP) losses, Dunleer and Corduff, followed by Timoleague and Ballycanew exhibited more susceptibility to changes in weather patterns manifested in the 11 year long time series (Table 3).

Table 2

Table 3

4. Discussion

Discharge, in relation to its fluctuations with weather changes, was one of the most significant drivers of regulating concentrations of both N and P. Investigation of the dynamics between discharge and nutrient concentrations under changing climate (and different hydrometeorological conditions) is gaining more interest (D’Amario et al., 2021; Pettersen et al., 2021). It is now understood that NO₃-N and TP concentration patterns in relation to discharge tend to vary for different geographical settings and be influenced by other drivers e.g., anthropogenic activities (D’Amario et al., 2021; Vystavna et al., 2023), or changes in weather (Ramos et al., 2022; Ezzati et al., 2023). Studies on the efficiency of management practices under climate change scenarios have shown that depending on the influential catchment characteristics, some watersheds would be more resilient in face of increased intensity of rainfall (Yuan and Koropecj-Cox, 2022). Extreme rainfall events and rainfall outside the growing season are known to not only influence agricultural production by prolonging the growing season, but they are also strongly correlated to discharge, N concentration (Øygarden et al., 2014), and P concentration (Ramos et al., 2022). For example a year with normal precipitation has shown strong chemodynamic behaviour in terms of N

concentration-discharge while a year with above normal precipitation suggested chemostatic behaviour (Miller and Lyon, 2021). Chemostatic behaviour is hydrologically controlled in the catchment (Pohle et al., 2021), hence, the concentrations remain stable over a wide range of flows in a concentration-discharge (c-q) pattern due to transport limitation (Bieroza et al., 2018). On the contrary, the rate of change in pollutant concentration in a chemodynamic c-q pattern, is larger than the flow change suggesting a source-limitation mechanism (Basu et al., 2010). Precipitation can change the quantity of discharge as well, which would consequently impact the nutrient fluxes (Grusson et al., 2021). This impact would show itself as either a sudden increase in concentrations due to washing out of diffuse sources of nutrients and sediments (Delkash et al., 2018), e.g. increasing the average P loads by 9% (Ockenden et al., 2016), or gradual decreases due to dilution by the water flow (Zhao et al., 2018).

Significant increases in Tair and Tsoil were observed in both Ballycanew and Castledockrell study catchments during summer months (June, July, August). These drivers suggested to have significant impacts on nutrient concentrations in all catchments. The impact of higher temperature on NO₃-N concentrations in diverse geographical locations has also been observed in other studies (Buciene et al., 2019; Ezzati et al., 2023). Higher temperatures (warmer summer or winter) will create large N pools in soils and increase mineralisation and nitrification rates (Wang et al., 2016) which leads to an increased risk of concentrations at the start of any ensuing wet period (Melander & Jordan, 2021). Rising temperatures would not only increase the risk of droughts and fluctuating quantities of discharge (Vystavna et al., 2023), but will also change soil shear strength (Zhang et al., 2022), which would consequently cause decreases in multi-year flow and mean active water storage capacity in catchments (Vystavna et al., 2023). On the other hand, water limitation in summer droughts affects grass growth by reducing uptake from spring N fertilizers. In order to compensate for that, higher mineral N fertilizers may in some cases be applied (Kundel et al., 2021) which would increase the risk of concentrations during winter time. Also, as a large pool of excessive N has been built up, a higher water discharge will increase losses with the first rain after a prolonged summer period (Ulen et al., 2009; Outram et al., 2014; Mellander and Jordan, 2021). However, if the groundwater has a high storage potential or long residence time (Fenton et al., 2011), or the ditches have low equilibrium phosphorus concentration (Ezzati et al., 2021), the excess N and P, respectively, will be naturally attenuated before leaving the catchment.

404

405 The three groundwater-fed catchments (Castledockrell, Cregduff, and Timoleague with
406 dominant soil types of Al-rich, Al and Ca rich, and Fe-rich, respectively) showed a chemostatic
407 behaviour for P which was not affected by changes in the weather patterns (Mellander et al.,
408 2022^b). Catchment characteristics can significantly change the relationships between discharge
409 and stream water nutrient concentrations by changing the strength and nature of stream flows
410 (Gao et al., 2021) or provide a natural attenuation function due to soil chemistry (Ezzati et al.,
411 2021). The biogeochemical properties of soils in these catchments are also indicative of soil
412 high affinity to bind P as long as the soil has not become P-saturated to act as a P-source rather
413 than a P-sink (Ezzati et al., 2020). In Castledockrell, the significant increase in NO₃-N
414 concentrations in April over the 11 years coincided with chemical fertiliser application and
415 ploughing/sowing (Fig. 2). This catchment is not hydrologically risky (i.e., extreme
416 hydrological events do not create a shift in nutrient dynamics), and was categorised as
417 chemostatic in terms of P (i.e., higher flow will not lead to higher nutrient losses (Bieroza et
418 al., 2018)). According to Mellander & Jordan (2021), the groundwater-fed well drained “N loss
419 risky” Castledockrell catchment, becomes “P loss risky” during winter storms. This transition
420 happens when SMD has been 0 mm for two or more consecutive days, and the effective rain
421 exceeded 15 mm/day (Mellander and Jordan, 2021). Timoleague catchment experienced a
422 significant ($P < 0.001$) increase in P-concentrations during September and a first-level
423 significant increase ($P < 0.1$) in NO₃-N concentrations in January, April, and November, over
424 the 11 year study period. Spring barley, wheat, and maize were the dominant crops, with the
425 latter requiring a large amount of slurry spreading which starts from mid-January and continues
426 until October (inclusive). Chemical fertilisers were applied in April. The maize fields are also
427 closer to the catchment outlet, which may contribute to the high concentrations in the outlet.

428

429 In the Ballycanew catchment, the increases in TP and TRP concentrations occurred during
430 January, May, and September; yet, no particular farming practice could be associated with
431 increases in P flux during January. Grazing (which occurs February-November) coincided with
432 cutting grasses in May. However, slurry spreading in September coincided with a significant
433 increase in the trend of monthly Tair values. Assessment of the high-resolution data revealed
434 that precipitation exceeding 10 mm per day consistently caused a sharp increase in P
435 concentrations. The elevated level of concentration returned to background values after few

days when the nutrient sources were exhausted or not hydrologically connected. In addition, continuous rainfall over several days caused an increase in TP concentrations, regardless of the amount of rainfall as more source areas were connected. Remarkably, an extreme hydrological event can turn the P-risky catchment of Ballycanew to N-risky as well. Ireland experienced a series of extreme weather events including a severe summer drought in 2018 (Falzoi et al., 2019). Following the rainfalls in September in the same year, the daily average concentration of NO₃-N reached about 6 times larger than the Environmental quality Standard (EQS) in Ballycanew and it remained high throughout the year (Mellander and Jordan, 2021).

Corduff exhibited the highest chemodynamic behaviour among the studied catchments. According to Mellander et al. (2018), the average annual runoff coefficients (fraction of rainfall that appears as runoff) between 2009-2015, in descending order, were Corduff (0.57), Timoleague (0.55), Castledockrell (0.54), Dunleer (0.48), and Ballycanew (0.48), (Cregduff is a spring contribution zone). This explains the high chemodynamic behaviour (Bieroza et al., 2018) of Corduff (highly affected by changes in the climate due to hydrological characteristics). The slurry spread during July seemed to have a significant impact on the concentrations of all nutrients. Corduff is mainly permanent grassland with sheep being the main grazing animal from March-November. Increasing trends in P (TP and TRP) and NO₃-N concentrations in Dunleer **were** observed in six months and one month, respectively. This increase in monthly losses occurred almost as frequently as in Corduff, during October-January, and March-April. However, potential evapotranspiration seemed to be the only driver with an increasing trend (i.e. less precipitation), which may lead to slurry spreading as the weather condition is good. According to local knowledge transfer expert, the ploughing/sowing in Dunleer take place during October and November, potatoes are harvested during November, and beets are harvested during November/December. Hence, further farm-scale investigation is required to understand the changes in nutrient concentrations leaving the Dunleer catchment outlet as it is the most complex catchment among all those included herein.

The results of analysing high temporal resolution historical weather and water quality data of the past 11 years suggested a relationship between increasing trends in climatic drivers and trends in within-stream nutrient concentrations. However, the significance of the correlation between nutrient concentrations and changes in weather patterns was defined by the catchment

characteristics including the drainage status and soil chemistry. The present study showed that the non-climatic indices, such as discharge and SS, which are also affected by the changing weather patterns, were significant drivers of nutrient concentrations mostly during the warmer time period of the study years. Hence, current mitigation strategies require explicit consideration of the more frequent occurrence of extreme weather events and their impacts on water pollutant concentrations. It is important to note that there may be inter-seasonal trends that are not identified when assessing inter-annual trends. For example, in the arable Castledockrell catchment, it is recommended to focus on improving N mitigation measures and/or changes in management decisions during April and November. However, in the other grassland catchments of Corduff and Timoleague, the summer period (May-August) plus October-November; and January, April, and November are important, respectively. Meanwhile, an improved focus on P mitigation and management is required in Ballycanew during January, May, and September; in Dunleer during March, July, October-January, and in Timoleague during September and December.

5. Conclusions

This study used long-term and high temporal resolution data (10-min) to investigate the N, P, and SS concentration trends in relation to the changing hydro-meteorological patterns in six agricultural catchments during 2010-2021. The results of historical data analysis showed that some of the catchments exhibited changes in the daily average trends of $\text{NO}_3\text{-N}$, TP, TRP, and/or SS concentrations with an associated significant increasing trend of climatic drivers, i.e. air temperature, soil temperature, and precipitation across the same months over the 11 years of monitoring period. The soil chemistry and drainage status of the study catchments are the non-climatically related drivers of nutrients concentrations while changes in air and soil temperature, SMD, and precipitation are the main climatic drivers regulating the nutrient concentrations. The extent of the impact of these climatic drivers highly depend on unique catchment characteristics. Discharge had a significant impact on the concentrations of $\text{NO}_3\text{-N}$, TP, TRP, and SS; air and soil temperature were significantly correlated to $\text{NO}_3\text{-N}$ losses; and precipitation was the major contributor to regulating TP and TRP concentrations across the different catchments.

The inter-seasonal and inter-annual trends of losses revealed that the climatically driven changes during the same month over the 11 years are manifested differently in different

catchments due to their individual hydroclimatological characteristics and management practices. Hence, catchment specific mitigation strategies are required to overcome future challenges to water quality management and the selection of climate-resilient mitigation measures in view of more frequently occurring extreme weather events and projected climate change scenarios.

Hence, building upon the results of the current study herein, we recommend that further studies are needed to investigate possible increases of nutrient concentrations into water bodies from agricultural catchments, using the current projected climate change scenarios in Ireland. This would eventually lead to developing future climate-resilient mitigation strategies.

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References

- Abbass, K., Qasim, M.Z., Song, H., Murshed, M., Mahmood, H., Younis, I., 2022. A review of the global climate change impacts, adaptation, and sustainable mitigation measures. *Environmental Science Pollution Research*, 29, 42539–42559
- Basu, N. B., Destouni, G., Jawitz, J. W., Thompson, S. E., Loukinova, N. V., Darracq, A., et al., 2010. Nutrient loads exported from managed catchments reveal emergent biogeochemical stationarity. *Geophysical Research Letters*, 37, 23, 1–n.
- Basu, N. B., Van Meter, K. J., Byrnes, D. K., Van Cappellen, P., Brouwer, R., Jacobsen, B. H., Jarsjö, J., Rudolph, D.L., Cunha, M.C., Nelson, N., Bhattacharya, R., Destouni, G., Olsen, S.B., 2022. Managing nitrogen legacies to accelerate water quality improvement. *Nature Geoscience*, 15, 2, 97–105
- Beusen, A. H. W., Doelman, J. C., Van Beek, L. P. H., Van Puijenbroek, P. J. T. M., Mogollón, J. M., Van Grinsven, H. J. M., Stehfest, E., Van Vuuren, D.P., Bouwman, A.F., 2022. Exploring river nitrogen and phosphorus loading and export to global coastal waters in the Shared Socio-economic pathways. *Global Environ. Change*, 72, 102426.
- Bieroza, M.Z., Heathwaite, A.L., Bechmann, M., Kyllmar, K., Jordan, P., 2018. The concentration-discharge slope as a tool for water quality management. *Science of the Total Environment*, 630, 738-749.
- Bučienė, A., Povilaitis, A., Langas, V., Bučas, M., Petkuvienė, J., Vaičiūtė, D., Gužys, S., 2019. Changes in Nutrient Concentrations of Two Streams in Western Lithuania with Focus on Shrinkage of Agriculture and Effect of Climate, Drainage Runoff and Soil Factors. *Water*. 11, 1590.
- Caretta, M. et al. , 2022. “Water”, in H.-O. Pörtner et al. (eds.), *Climate Change 2022: Impacts, Adaptation, and Vulnerability. Contribution of Working Group II to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change*, Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.

554

555 Coffey, R., Paul, M., Stamp, J., Hamilton, A., Johnson, T., 2018. A review of water quality
556 responses to air temperature and precipitation changes 2: Nutrients, algal blooms, sediment,
557 pathogens. *Journal of American Water Resource Association*, 55, 4, 844-868.

558

559 Covino, T.P., Wlostowski, A.N., Gooseff, M.N., Wollheim, W.M., Bowden, W.B., 2021. The
560 seasonality of in-stream nutrient concentrations and uptake in Arctic headwater streams in the
561 northern foothills of Alaska's Brooks range. *JGR Biogeosciences*, 126, e2020JG005949.

562

563 D'Amario, S.C., Wilson, H.F., Xenopoulos, M.A., 2021. Concentration-discharge relationships
564 derived from a larger regional dataset as a tool for watershed management. *Ecological
565 Applications*, 31, 8, e02447.

566

567 DAFM, 2017. Nitrate explanatory handbook for good agricultural practice for the protection
568 of waters regulation 2018. Department of Agriculture Fisheries and the Marine, Ireland.

569

570 Ehrhardt, S., Ebeling, P., Dupas, R., Kumar, R., Fleckenstein, J. H., and Musolff, A., 2021.
571 Nitrate transport and retention in western European catchments are shaped by hydroclimate
572 and subsurface properties. *Water Resources Research*, 57, 10, e2020WR029469.

573

574 EPA, 2021. Water quality in Ireland 2016-2021. EPA, Johnstown Castle, Wexford, Ireland.
575 Retrieved from [https://www.epa.ie/publications/monitoring--assessment/freshwater--
576 marine/EPA_WaterQualityReport2016_2021.pdf](https://www.epa.ie/publications/monitoring--assessment/freshwater--marine/EPA_WaterQualityReport2016_2021.pdf)

577

578 EPA, (2021a) Research 386: Climate Status Report for Ireland 2020. Report prepared by
579 Walther CA, García C, Dwyer N (eds) EPA, Wexford, Ireland. ISBN: 978-1-80009-009-5.
580 Available at: [https://www.epa.ie/publications/research/climate-change/research-386-the-
581 status-of-irelands-climate-2020.php](https://www.epa.ie/publications/research/climate-change/research-386-the-status-of-irelands-climate-2020.php)

582

583 EPA, 2023. Water Quality in 2022: An Indicators Report. EPA, Johnstown Castle, Wexford,
584 Ireland. Retrieved from [https://www.epa.ie/publications/monitoring--assessment/freshwater--
585 marine/Water-Quality-2022-Indicator-Report-Web.pdf](https://www.epa.ie/publications/monitoring--assessment/freshwater--marine/Water-Quality-2022-Indicator-Report-Web.pdf)

586

587 Ezzati, G., Fenton, O., Healy, M.G., Christianson, L., Feyereisen, G.W., Thornton, S., Chen,
588 Q., Fan, B., Ding, J., Daly, K., 2020. Impact of P inputs on source-sink P dynamics of sediment
589 along an agricultural ditch network. *Journal of Environmental Management*, 257, 109988.

590

591 Ezzati, G., Kyllmar, K., Barron, J., 2023. Long-term water quality monitoring in agricultural
592 catchments in Sweden: Impact of climatic drivers on diffuse nutrient loads. *Science of the Total*
593 *Environment*, 8864, 160978.

594

595 Falzoi, S., Gleeson, E., Lambkin, K., Zimmermann, J., Marwaha, R., O'Hara, R., Green, S.,
596 Fratianni, S., 2019. Analysis of the severe drought in Ireland in 2018. *Weather* 74, 368–373.

597

598 Fenton, O., Schulte, R.P.O., Jordan, P., Lalor, S.T.J., Richards, K. G., 2011. Time lag: a
599 methodology for the estimation of vertical and horizontal travel and flushing timescales to
600 nitrate threshold concentration in Irish aquifers. *Environmental Science & Policy*, 14, 419–431.

601

602 Gao, C., Li, W., Cui, L., MA, Q., Cai, J., 2021. Catchment landscape components alter
603 relationships between discharge and stream water nutrient ratios in the Xitiao River Basin
604 China. *Scientific Report*, 11, 10466.

605

606 Gascuel-Oudou, C., Fovet, O., Faucheux, M., Salmon-Monviola, J., Strohmenger, L., 2023.
607 How to assess water quality change in temperate headwater catchments of western Europe
608 under climate change: examples and perspectives. *Comptes Rendus. Géoscience*, 1-11.

609

610 Grusson, Y., Wesstrom, I., Svedberg, E., Joel, A., 2021. Influence of climate change on water
611 partitioning in agricultural watersheds: examples from Sweden. *Agricultural Water*
612 *Management*, 249, 106766.

613

614 Hadush, M., Conor, M., 2023. Climate change impacts on the frequency, magnitude and
615 duration of meteorological droughts for the island of Ireland using SPI and SPEI. *Internaitonal*
616 *Journal of Climatology*, JOC-23-0033

617

Hughes, J., Cowper-Heays, K., Oleasson, E., Bell, R., Stroombergen, A., 2021. Impacts and implications of climate change on wastewater systems: A New Zealand perspective. *Climate risk Management*, 31, 100262.

Jordan, Y.C., Ghulam, A, Hartling, S., 2014. Traits of Surface Water Pollution under Climate and Land Use Changes: A Remote Sensing and Hydrological Modeling Approach. *Earth-Science Reviews*, 128, 181–195.

Kendall, M.G., 1948. Rank correlation methods. Griffin.

Kundel, D., Lori, M., Fließbach, A.; van Kleunen, M., Meyer, S., Mäder, P., 2021. Drought Effects on Nitrogen Provisioning in Different Agricultural Systems: Insights Gained and Lessons Learned from a Field Experiment. *Nitrogen*, 2, 1-17.

Lintern, A., Liu, S. C., Minaudo, C., Dupas, R., Guo, D. L., Bende-Michl, U., and Duvert, C., 2021. The influence of climate on water chemistry states and dynamics in rivers across Australia. *Hydrological Processes*, 35, 12, e14423.

Lungarska, A., Chakir, R., 2018. Climate-induced land use change in France: Impacts of agricultural adaptation and climate change mitigation. *Ecological Economics*, 147, 134–154.

Maúre, E. d., R., Terauchi, G., Ishizaka, J., Clinton, N. and deWitt, M., 2021. Globally consistent assessment of coastal eutrophication. *Nature Communications*, 12, 6142.

Melland, A.R., Mellander, P.-E., Murphy, P., Wall, D.P., Mechan, S., Shine, O., Shortle, G., Jordan, P., 2012. Stream water quality in intensive cereal cropping catchments with regulated nutrient management. *Environmental Science & Policy*, 24, 58–70.

Mellander, P.-E., Galloway, J., Hawtree, D., Jordan, P., 2022^a. Phosphorus mobilization and delivery estimated from long-term high frequency water quality and discharge data. *Frontiers in Water*, 4, 917813.

Mellander, P. -E., Jordan, P., Melland, A. R., Murphy, P. N. C., Wall, D. P., Mehan, S., et al., 2013. Quantification of phosphorus transport from a karstic agricultural watershed to emerging spring. *Environmental Science Technology*, 47, 6111–6119.

Mellander, P.-E., Lynch, M.B., Galloway, J., Zurovec, O., McCormack, M., O’Neil, M., Hawtree, D., Burgess, E., 2022^b. Benchmarking a decade of holistic agro-environmental studies within the Agricultural Catchments Programme. *Irish Journal of Agricultural and Food Research*, 1-17.

Mellander, P.-E., Melland, A. R., Jordan, P., Wall, D. P., Murphy, P. N. C., and Shortle, G., 2012. Quantifying nutrient transfer pathways in agricultural catchments using high temporal resolution data. *Environmental Science &Policy*, 24, 44–57.

Mellander, P.-E., Jordan, P., 2021. Charting a perfect storm of water quality pressures. *Science of the Total Environment*, 787, 147576.

Mellander, P.-E., Jordan, P., Bechmann, M., Shore, M., McDonald, N.T., Fovet, O., Gascuel-Odoux, C., 2018. Integrated climate-chemical indicators of diffuse pollution. *Natural Sciences Report*, 8, 1, 944.

Mellander, P.-E., Jordan, P., Shore, M., Melland, A.R., Shortle, G., 2015. Flow paths and phosphorus transfer pathways in two agricultural streams with contrasting flow controls. *Hydrological Processes*, 29, 3504–3518.

Meresa, H., Murphy, C., 2023. Climate change impacts on spatial and seasonal meteorological drought characteristics for the island of Ireland. *International Journal of Climatology*, 1-17.

Miller, S.A., Lyon, S.W., 2021. Tile drainage increases total runoff and phosphorus export during wet years in the Western Lake Erie Basin. *Frontiers in Water*, 3.

Moal, M.L., Gascuel-Odoux, C., Ménesguen, A., Souchon, Y., Étrillard, C., Levain, A., Moatar, F., Pannard, A., Souchu, P., Lefebvre, A., Pinay, G., 2019. Eutrophication: a new wine in an old bottle?, *Science of the Total Environment*, 651, 1–11.

Murphy, R.R., Perry, E., Harcum, J., Keisman, J., 2019. A Generalized Additive Model approach to evaluating water quality: Chesapeake Bay case study. *Environmental Modelling & Software*, 118, 1-13.

Outram, F.N., Lloyd, C.E.M., Jonczyk, J., Benskin, C. McW. H., Grant, F., Perks, M.T., Deasy, C., Burke, S.P., Collins, A.L., Freer, J., Haygarth, P.M., Hiscock, K.M., Johnes, P.J. and. Lovett, A.A., 2014. High frequency monitoring of nitrogen and phosphorus response in three rural catchments to the end of the 2011-2012 drought in England. *Hydrology and Earth System Sciences* 18, 3429-3488.

Øygarden, L., Deelstra, J., Lagzdins, a., Bechmann, M., Greipsland, L., Kyllmar, K., Povilaitis, A., Iital, A., 2014. Climate change and the potential effects on runoff and nitrogen loads in the Nordic Baltic region. *Agricultural Ecosystem environment*, 198, 114-126.

Paul, M.P., Stamp, J., Hamilton, A., Coffey, R., Johnson, T., 2019. A Review of Potential Climate Change Effects on U.S. Water Quality - Part I: Sea Level Rise, Flow, and Temperature. *Journal of the American Water Resources Association*, 55, 4, 824–843.

Pettersen, R.J., Blicher-Mathiesen, G., Rolighed, J., Andersen, H.E., Kronvang, B., 2021. Three decades of regulation of agricultural nitrogen loads experiences from the dansih agricultural monitpring program. *Science of the Total Environment*, 787, 147619.

Pohle, I., Bagaley, N., Palarea-Albaladejo, J., Stutter, M., Glendell, M., 2021. A framework for assessing concentration-discharge catchment behaviour from low-frequency water quality data. *Water Recourses Research*, 57, e2021WR029692.

Ramos, M.C., Lizaga, I., Gaspar, L., Navas, A., 2022. The impacts of exceptional rainfall on phosphorus mobilisation in a mountain agroforestry catchment (NE, Spain). *CATENA*, 216, Part B, 106407.

Refsgaard, J.C., Arnbjerg-Nielsen, K., Drews, M., Halsnæs, K., Jeppesen, E., Madsen, H., Markandya, A., Olesen, J.E., Porter, J.R., Christensen, J.H., 2013. The role of uncertainty in climate change adaptation strategies—A Danish water management example. *Mitigation and Adaptation Strategies for Global Change*, 18, 337–359.

Samaniego, L., Thober, S., Kumar, R., Wanders, N., Rakovec, O., Pan, M., Zink, M., Sheffield, J., Wood, E.F. and Marx, A., 2018. Anthropogenic warming exacerbates European soil moisture droughts. *Nature Climate Change*, 8, 421-426

Sandström, S., Futter, M.N., Kyllmar, K., Bishop, K., O’Connell, D.W., Djodjic, F., 2020. Particulate phosphorus and suspended solids losses from small agricultural catchments: Links to stream and catchment characteristics. *Science of the Total Environment*, 711, 134616.

Sen, P., 1968. Estimates of the regression coefficient based on Kendall's Tau. *Journal of American statistician Association*, 63, 324.

Seyedhashemi, H., Vidal, J.-P., Diamond, J.S., Thiéry, D., Monteil, C., Hendrickx, F., Maire, A., Moatar, F., 2022. Regional, multi-decadal analysis on the Loire River basin reveals that stream temperature increases faster than air temperature. *Hydrology and Earth System Sciences*, 26, 9, 2583–2603.

Schulte, R.P.O., Diamond, J., Finkel, K., Holden, N.M., Brereton, A.J., 2005. Predicting the soil moisture conditions of Irish grasslands. *Irish Journal of Agricultural and Food Research*, 44, 95-110

Sherriff, S.C., Rowan, J.S., Fenton, O., Jordan, P., Melland, A.R., Mellader, P-E., O hUallachain, D., 2016. Storm event suspended sediment discharge hysteresis and controls in agricultural watersheds: Implications for watershed scale sediment management. *Environmental Science & Technology*, 50, 1769-1778.

Strohmenger, L., Fovet, O., Akkal-Corfini, N., Dupas, R., Durand, P., Faucheux, M., Gruau, G., Hamon, Y., Jaffrezic, A., Minaudo, C., Petitjean, P., Gascuel-Oudou, C., 2020. Multitemporal relationships between the hydroclimate and exports of carbon, nitrogen, and phosphorus in a small agricultural watershed. *Water Resources Research*, 56, 7, e2019WR026323.

Sulis, M., Paniconi, C., Rivard, C., Harvey, R., Chaumont, D., 2011. Assessment of climate change impacts at the catchment scale with a detailed hydrological model of surface-subsurface interactions and comparison with a land surface model. *Water Resources Research*, 47, 1.

Ulén, B., Johansson, G., 2009. Long-term nutrient leaching from a Swedish arable field with intensified crop production against a background of climate change. *Acta Agriculturae Scandinavica*, 59, 157–169.

von Brömssen, C., Betnér, S., Fölster, J., Eklöf, K., 2021. A toolbox for visualizing trends in large-scale environmental data. *Environmental Modelling & Software*, 136.

Vystavna, Y., Paule-Mercado, M.C., Schmidt, S.I., Hejzlar, J., Porcal, P., Matiatos, I., 2023. Nutrient dynamics in temperate European catchments of different land use under changing climate. *Journal of Hydrology: Regional Studies*, 45, 101288.

Warner, W., Zeman-Kuhnert, S., Heim, C., Nachtigall, S., Licha, T., 2021. Seasonal and spatial dynamics of selected pesticides and nutrients in a small lake catchment – Implications for agile monitoring strategies. *Chemosphere*, 281, 130736.

Weng, H., Bechmann, M., Krogstad, T., Skarbøvik, E., 2020. Climate effects on land management and stream nitrogen concentrations in small agricultural catchments in Norway. *Ambio* 49, 1747–1758.

Winter, C., Nguyen, T.V., Musolff, A., Lutz, S.R., Rode, M., Kumar, R., Fleckenstein, J.H., 2023. Droughts can reduce the nitrogen retention capacity of catchments. *Hydrology and Earth System Sciences*, 27, 303-318.

- Wood, S. N., 2006. Low rank scale invariant tensor product smooths for generalized additive mixed models. *Biometrics*.
- Wood, S.N., 2017. *Generalized Additive Models: An Introduction with R*, Second Edition (2nd ed.). Chapman and Hall/CRC.
- Yang, X., Jomaa, S., Zink, M., Fleckenstein, J. H., Borchardt, D., Rode, M., 2018. A New Fully Distributed Model of Nitrate Transport and Removal at Catchment Scale. *Water Resources Research*, 54, 5856–5877.
- Yuan, Y., Koropecj-Cox, L., 2022. SWAT model application for evaluating agricultural conservation practice effectiveness in reducing phosphorous loss from the Western Lake Erie Basin. *Journal of Environmental Management*, 302(Pt A), 114000.
- Yue, S. Pilon, P., Cavadias, G., 2002. Power of the mann-kendall and Spearman's rho tests for detecting monotonic trends in hydrological series. *Journal of Hydrology*, 259, 254-271.
- Zhao, Z., Liu, G., Liu, Q., Huang, C., Li, H., 2018. Studies on the Spatiotemporal Variability of River Water Quality and Its Relationships with Soil and Precipitation: A Case Study of the Mun River Basin in Thailand. *International Journal of Environmental Research and Public Health*, 15, 2466.
- Zhang, S., Liu, Y., Du, M., Shou, G., Wang, Z., Xu, G., 2022. Nitrogen as a regulator for flowering time in plant. *Plant and Soil*, 480, 1–29.
- Zhang, Y., Zhao, D., Zheng, Q., Huang, Y., Jiang, F., Wang, M-K., Lin, J., Huang, Y., 2022. Evaluating the effects of temperature on soil hydraulic and mechanical properties in the collapsing gully areas of south China. *CATENA*, 218, 106549.

Caption of Figures

Figure 1. Location of the six catchments monitored by the Agricultural Catchments Programme. The time series plots show daily average nitrate-N ($\text{NO}_3\text{-N}$) and total phosphorus (TP) concentrations (mg L^{-1}) for the time period of 2010-2021. The values on the y-axis represent maximum concentration detected in the catchments.

Figure 2. Summary of the typical farming calendar for each of the six catchments monitored by the Irish Agricultural Catchments Programme. In order to assist visualisation, similar bar colour are used for similar activity across different catchments.

Figure 3. Box and whisker plots of daily concentrations of SS, $\text{NO}_3\text{-N}$, TP and TRP, and the average annual total discharge leaving the catchment outlets for the period 2010– 2021. Cregduff is a karst spring contribution zone and therefore does not have SS data. The line in the boxplots marks the median point of the data. Whiskers show the upper and lower quartiles (75th and 25th percentiles, respectively). Outliers are shown as circles. The dotted horizontal line shows environmental quality standards (EQS) for $\text{NO}_3\text{-N}$ and TRP.

Figure 4. The average monthly trend analysis of $\text{NO}_3\text{-N}$, TP, TRP concentrations, and explanatory variables during 2010-2021 using Mann-Kendall tests. The red colour indicates increasing trends and the green colour indicates decreasing trends. The level of significance is presented by the different shades of any individual colour.

Caption of Tables

Table 1. Catchments characteristics. Precipitation, soil and air temperatures, potential evapotranspiration (PET), and soil moisture deficit (SMD) values are averaged for the 11 hydrological years (1st October 2010 – 30th September 2021).

Table 2. The significance of climatic drivers in regulating NO₃-N concentrations with a significant monthly trend during water years spanning 2010-2021. Only catchments and months with changes are included.

Table 3. The significance of climatic drivers in regulating TP and/or TRP concentrations with a significant monthly trend during water years spanning 2010-2021. . Only catchments and months with changes are included.