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1 **Impacts of changing weather patterns on the dynamics of water pollutants in**
2 **agricultural catchments: Insights from 11-year high temporal resolution data analysis**

3
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10
11 **Abstract:**

12 Widespread and long-term shifts in weather patterns are contributing to further degradation of
13 surface water quality. This challenge caused by the increasing frequency of extreme weather
14 events requires appropriate adaptation of current mitigation strategies. But to confirm the need
15 to redesign such strategies, an understanding of the impacts of increasing weather extremes on
16 pollutant losses in different catchment types is required. With this in view, this study
17 investigated the impact of changing weather patterns on the inter-seasonal and inter-annual
18 dynamics of nutrient losses in six agricultural catchments in Ireland over 11 years. The high
19 temporal resolution data (10-min) from these intensively managed catchments represented
20 different characteristics and management practices. Mann-Kendall Trend Analysis and
21 Generalised Additive Models were used to study nutrient concentration trends, and to
22 investigate the significance of water discharge, precipitation, potential evapotranspiration, soil
23 moisture deficit, air temperature, and soil temperature on the losses of nutrients, respectively.

24
25 The analysis of historical data revealed changes in the trends of daily average nitrate (NO₃-N),
26 phosphorus (P), and suspended sediment (SS) concentrations in association with significant
27 increasing trends in air temperature, soil temperature, and precipitation across the same month
28 over 11 years of monitoring. While discharge was significantly contributing to the
29 concentrations of NO₃-N, P, and SS across different catchments, air and soil temperature were
30 significantly correlated to NO₃-N losses, and precipitation was the major contributor to
31 regulating P (total P and total reactive P) concentrations. In short, air temperature, soil

32 temperature, soil moisture deficit, and precipitation were the main climatic drivers regulating
33 the nutrient concentrations while the soil chemistry and drainage status were the non-
34 climatically related drivers.

35

36 The results revealed that the extent of the impact of climatic drivers depends on catchment
37 characteristics. Therefore, expanding the application of this type of study would facilitate better
38 understanding of current and future challenges to water management and provision of climate-
39 resilient mitigation strategies for different catchment typologies.

40

41

42

43 **Keywords:**

44 Nitrate, phosphorus, suspended solids, climate change, farming, water quality

45

46

47 **1. Introduction**

48 Excess nutrients (i.e. nitrogen (N) and phosphorus (P)) and suspended sediments coming from
49 agriculture can be delivered into lakes, streams, estuaries, and coastal waters causing
50 degradation of water quality (Maúre et al., 2021; Basu et al., 2022; Beusen et al., 2022). Current
51 European Legislation requires the application of mitigation strategies to meet the goal of the
52 Water Framework Directive (WFD) (2000/60/EC) to achieve ‘good ecological status’ in all
53 waterbodies by 2027. However, the overall ecological status of waterbodies has declined in
54 some countries, including Ireland, which has reported increasing nutrient concentrations,
55 mainly from agricultural sources, over the period 2016-2021 (EPA, 2021^a). Hence, a key
56 challenge is the need to balance agricultural intensification for food security with achieving
57 ‘good status’ in waterbodies (Moal et al., 2019). This is a global and growing issue which also
58 impacts economic/social welfare as well as the sustainability of ecosystems and biodiversity
59 (Weng et al., 2020).

60

61 To achieve an improvement in waterbody status, an understanding of both the nutrient sources
62 (e.g., soils and farming practices), and climate drivers of water pollutant mobilisation/delivery

63 are required. Recent research has particularly highlighted the need for a comprehensive
64 analysis of the links between the weather patterns associated with changing climate and river
65 pollutant concentrations (Mellander and Jordan, 2021). Climate change is exerting extra
66 pressures on water quality as an increasing frequency of extreme weather events (drought and
67 flood) is being observed (Gascuel-Oudoux et al., 2022). For example, reduced discharge during
68 drought conditions can lead to the reduced dilution of pollutant discharges from point-sources
69 and therefore can increase fluvial nutrient concentrations (Hughes et al., 2021). Changes in soil
70 and water temperature and soil water content can also modify water quality indirectly
71 (Seyyedhashemi et al., 2022). Drought can alter hydrological connectivity (Erhardt et al., 2021)
72 and disconnect shallow flow pathways from the river network (Yang et al., 2018). Additionally,
73 during the first heavy rain or so-called ‘first flush’ after drought conditions decreased
74 denitrification and the subsequent flushing of accumulated nutrients have been observed
75 (Outram et al., 2014; Strohmenger et al., 2020; Zhang et al., 2022; Winter et al., 2023).

76

77 According to the Irish Environmental Protection Agency (EPA, 2021^b), starting from 1980s,
78 every decade has been warmer than the previous one and during the last ten-year decade (2010-
79 2019), the average temperature was the warmest on record. An increase in the magnitude,
80 frequency and duration of droughts is further expected during summer times (Merese and
81 Murphy, 2023). While the average annual precipitation in Ireland has increased approximately
82 5% (60mm) during 1981-2010 compared to 1961-1900, substantial decreases in average annual
83 spring and summer precipitation, and significant increase in frequency of extreme events in
84 winter and autumn are expected (EPA, 2021^b). Weather extremes have been shown to have
85 diverse impacts on nutrient loads within different agricultural landscapes depending upon their
86 specific characteristics (Mellander et al., 2021; Ezzati et al., 2023) and boundary conditions
87 (catchment boundaries) (Ehrhardt et al., 2021). These impacts generally correlate with changes
88 in runoff volume (Coffey et al., 2021) and are therefore strongly controlled by precipitation
89 and air temperature (Jordan et al., 2014; Paul et al., 2019; Ezzati et al., 2023; Hadush &
90 Murphy, 2023). On the other hand, catchments, in a global scale, also have contrasting
91 characteristics such as soil type and texture (including average clay content), land use (the
92 fraction of arable land), management, and hydrology (Sandström et al., 2020) which result in
93 different responses to the same weather events. The differences between catchments are
94 comprised of variations in the sensitivity of hydrological responses to climatic drivers (Sulis et
95 al., 2011), and thus, variations in intrinsic controls relevant to nutrient mobilisation and

96 delivery and differences in farming systems. Hence, nutrient dynamics in catchments, despite
97 close geographic proximity but with different controls and mitigation strategies, have shown
98 to respond differently to the same large-scale weather extremes (Mellander & Jordan, 2021).

99

100 Existing management efforts and mitigation strategies have not only failed to improve water
101 quality in Ireland (EPA, 2023), but are also incapable of buffering the effects of weather
102 extremes in the already-vulnerable agricultural sector at a global scale. This failure is of greater
103 importance in countries where the agri-food sector is an integral part of the economy and
104 society (Abbass et al., 2022). According to Samaniego et al. (2018), Europe may face a 40
105 (± 24) % increase in drought affected areas in the absence of effective mitigation strategies
106 during the 21st century (Caretta et al., 2022). Therefore, when considering the more frequently
107 occurring extreme weather events, more targeted approaches and resilient management
108 strategies are required to deliver catchment-specific and climate-resilient adaptation. However,
109 uncertainty remains regarding the causes and severity of water quality degradation due to
110 changing climate (Refsgaard et al., 2013).

111

112 The modelling of water quality, which is often used to guide policymakers in developing
113 appropriate mitigation strategies, is facing new challenges in representing nutrient trends,
114 legacies, delivery, and mobilization in view of current and projected climate change scenarios
115 (Mellander et al., 2018; Mellander et al., 2022^b). The impacts of the changing weather patterns
116 have also been less considered in policy reviews (Mellander et al., 2018). According to
117 Gascuel-Oudoux et al. (2022), these challenges are due to three main reasons: 1. Lack of long
118 time series data for water quality and chemical concentrations compared to hydrologic fluxes
119 and storage measurements; 2. Lack of detailed understanding of climate-water quality
120 relationships compared to well-developed, but highly generalised, climate-hydrology models
121 at global and regional scales (Lintern et al., 2021); 3. Difficulty in distinguishing climate effects
122 from those due to changes in farming activities (e.g. types of the crops) (Lungarska and Chakir,
123 2018). In addition, seasonality drives changes in nutrient concentrations as a consequence of
124 changes in hydrologic flow pathways, climate change, and associated biological processes
125 (Covino et al., 2021). However, understanding the seasonal dynamics of nutrients is highly
126 complex and ensuring good water quality requires catchment-specific approaches (Warner et
127 al., 2021).

128 Changes in **the** nutrient concentrations are influenced by mobilisation and hydrological
129 processes, driven by weather and agricultural management, and controlled by physical
130 catchments characteristics (e.g. soil/bedrock drainage and chemistry) and can as such be subtle
131 (Mellander et al., 2018). While nutrient concentrations can vary within any span of time, the
132 impacts of extreme runoff events are typically more drastic and can be exhibited immediately.
133 Changes in the nutrient loads, on the other hand, represent the cumulative quantity without
134 capturing fluctuations, while hydrology may also override any processes in mobilisation
135 (Mellander et al., 2022^a).

136 This study used 11-years (2010-2021) of sub-hourly (10-min basis) water quality (i.e.,
137 concentration data) and climatic data from six hydrologically contrasting agricultural
138 catchments in Ireland. The objectives were to statistically assess the presence of seasonal and
139 annual trends in nutrient and sediment concentrations using Mann-Kendall Trend Analysis,
140 and to evaluate the impact of climatic data (i.e. precipitation, air/soil temperature, soil moisture
141 deficit), on nutrient source loading and in view of management practices, using Generalised
142 Additive Models. Such analysis may facilitate a better understanding to counter future
143 challenges for water management and provision of climate- smart mitigation strategies.

144

145 **2. Materials and Methods**

146 In order to identify any existing trend in nutrient concentrations in view of recent extreme
147 weather events in Ireland (prolonged draught periods and/or heavy precipitation), and in order
148 to understand the importance of climatic variables in regulating nutrient losses throughout
149 different months of the year, long-term high-resolution water quality data from agricultural
150 catchments with contrasting physical and management factors were studied using different
151 statistical modelling techniques.

152

153 **2.1. Site study description**

154 Six agriculturally-dominated catchments (one karst spring contribution zone and five river
155 catchments) monitored by the Irish Agricultural Catchments Programme (Mellander et al.,
156 2022^{a,b}) were investigated (Fig. 1 and Tab. 1). The catchments are all intensive agriculture
157 areas and comprise a varying mix of grassland (pasture) and arable land use. They all have
158 different physical characteristics, i.e., slopes, soil types, drainage, and areas ranging from 3 to

159 31 km². The base-flow (Bf_i) indices and detailed fertiliser applications rates have been assessed
160 in Mellander et al., 2022^a.

161

162

Figure 1

163

164 Ballycanew (11.9 km²-Co. Wexford) is in the southeast of Ireland. This catchment has mostly
165 poorly drained soil with land use dominated by grasslands (77%) grazed by dairy cows. There
166 was a 30% increase in stocking rate (livestock per area) between 2010-2018. Grazing intensity
167 is highest in May and fertilisers are mostly N-based. December and January are a “closed
168 period” as defined by Nitrate Action Programme in Ireland (NAP) (DAFM, 2018), during
169 which grazing and the spreading of fertilisers are prohibited. Excess slurry stored in tanks is
170 typically spread on silage fields at the end of March. The hydrology is flashy (base-flow index
171 (Bf_i) of 0.63) with dominating surface/near surface pathways (Mellander et al., 2015).

172 Castledockrell (11.2 km²-Co. Wexford) is also located in southeast Ireland. This catchment has
173 mostly well-drained soils with 72% of the catchment under arable land use (66% of the total
174 area is under tillage which is the highest of the catchments studied). Hydrological pathways
175 are primarily subsurface (Mellander et al., 2016) with Bf_i of 0.78.

176 Corduff (3.3 km²-Co. Monaghan) is in northeast Ireland. This poorly drained catchment has no
177 tillage, has low chemical fertiliser application rates (as detailed in Mellander et al., 2022^a), and
178 is dominated by grassland for sheep and suckler cows. The catchment has a flashy hydrology
179 (Bf_i of 0.57) dominated by surface/near surface pathways.

180 Cregduff (31.2 km²-Co. Mayo) is a mainly permanent grassland karst spring contribution zone,
181 located in west Ireland. Sheep are primarily grazed and have a longer grazing period than cattle.
182 The soil is well drained and relatively thin. Hydrological pathways are dominantly subsurface
183 (Mellander et al., 2013) and Bf_i is 0.82. The rate of N-application in this catchment is low
184 relative to other catchments.

185 Dunleer (9.5 km²-Co. Louth) is in northeast Ireland. Farm types in this catchment are mixed at
186 approximately 40% grassland, half of which is grazed by dairy and half by beef cattle. This
187 catchment has moderately drained soils with flashy hydrology (Bf_i of 0.66) and mostly
188 surface/near surface pathways (Mellander et al., 2012). Chemical N fertilisers are spread in

189 April/March but available soil P has been notably increasing during the recent years. The
190 ploughing/sowing take place during October and November.

191 Timoleague (7.6 km²-Co. Cork) is in southwest Ireland, is well-drained and dominated by
192 grassland and subsurface hydrological pathways (Mellander et al., 2016) with Bf_i of 0.73.
193 There has been an increase in livestock grazing intensity and the percentage of land under
194 derogation (allowance to farm at livestock above the limit of 170 kg N ha⁻¹ under the EU
195 Nitrates Directive). The P-index (soil fertility) is also increasing more than expected. Maize is
196 one of the dominant crops which requires substantial slurry spreading typically between mid-
197 January and October.

198

199 The management practices for a typical farming calendar in each catchment were collated from
200 discussion with the local farm advisor and knowledge transfer expert (Fig. 2).

201

202

Figure 2

203

204 2.1.2. Hydrochemistry data collection

205 Bankside P analysers (Hach-Lange Sigmatax-Phosphax) (Melland et al., 2012) were located at
206 the catchments outlets which measure total digested P (TP) and total reactive phosphorus (TRP)
207 concentrations on unfiltered samples. The measuring range is 0.010 mg L⁻¹ to 5.000 mg L⁻¹,
208 and the detection limit is 0.010 mg L⁻¹. Total oxidized N (TON) was monitored using Hach-
209 Lange Nitratax SC-Plus UV instruments (co-located at the outlets). The measuring range is
210 0.1–50 mg L⁻¹ and assuming a low NO₂-N concentration (Melland et al., 2012), TON is
211 considered equivalent to NO₃-N. The measuring of all nutrients was carried out based on a 10-
212 min basis.

213 Stage-discharge rating curves on Corbett flat-v non-standard weirs have been established at the
214 individual catchment outlets. The ratings are based on the velocity-area method with an OTT
215 Acoustic Doppler Current meter (in WISKI-SKED software). An OTT Orpheus Mini vented
216 pressure instrument installed in a stilling well adjacent to the weirs records the water level
217 every 10 min. The river discharge is calculated by converting the water level to flow. In
218 Cregduff, the discharge is calculated using an ultrasonic sensor (Thermo-Fisher time-of-flight
219 area velocity) which is placed in an engineered uniform cross-section.

220 Suspended sediment (SS) concentrations were estimated using turbidity reading and
221 corresponding concentration-turbidity conversion relationships (excluding the Cregduff
222 catchment given its karst geology) (Sherriff et al., 2016).

223

224 **2.1.3. Weather data collection**

225 Precipitation, air temperature (T_{air}), soil temperature (T_{soil}), relative air humidity, solar
226 radiation, wind speed and wind direction are measured at ten-minute intervals by a weather
227 station (BWS200, Campbell Scientific, www.acpmet.ie) located in the central valley floor
228 within each study catchment. The effective rainfall was calculated by subtracting potential
229 evapotranspiration (derived from the Penman-Monteith equation (Monteith, 1965)) from
230 measured rainfall.

231 The soil moisture deficit (SMD), which responds to changes in air temperature and rainfall,
232 was calculated as the amount of rain needed to bring the soil back to field capacity using a
233 SMD model (Schulte et al., 2005).

234

235 **Table 1**

236

237 **2.3. Data analysis**

238 **2.3.1. Monthly trend analysis**

239 To investigate the seasonality that may be hidden in inter-annual trends, and compare the
240 nutrient concentrations dynamics in each month across different years, the temporal trends in
241 average daily values of climatic-related explanatory drivers and nutrient concentrations were
242 calculated using the non-parametric rank-based Mann-Kendall test (Kendall, 1975). The
243 analysis was carried out over 11 years of the monitoring period. This test accounts for non-
244 normality of hydrological (Yue et al., 2002) and climatological data (Partal and Kahya, 2006).
245 Hence, the method provided accurate estimation and enabled comparison of any increasing or
246 decreasing trends of variables in each particular month over more than a decade. The null
247 hypothesis assumed that the data (x_1, \dots, x_n) consist of n independent and identically distributed
248 random variables and $P < 0.05$ was considered as a significant trend for calculating slopes of
249 time series (Sen, 1968). The intercept (α) was calculated as followed:

250 $\alpha = x_{0.5} - s * y_{0.5}$ Eq.

251 1

252

253 where $x_{0.5}$ and $y_{0.5}$ are median values of variables.

254

255 **2.3.2. GAM analysis of significance of drivers**

256 Generalised Additive Models (GAMs) (Hastie and Tibshirani, 1990) were fitted to the daily
 257 average time series data, in order to evaluate the significance of the impact of different climate
 258 variables on regulating water pollutant losses over different months of the year (on condition
 259 that an increasing/decreasing trend existed, see section 2.3.1). GAMs, as an extension to
 260 generalized linear models with smoothing functions, are becoming a widely applied statistical
 261 test for evaluation of changes in water quality at large scales (Murphy et al., 2019) with multiple
 262 variables (Wood, 2006) and consider uncertainty and random effects of both parametric and
 263 non-parametric variables. Hence, GAMs add flexibility to the modelling of complex non-linear
 264 relationships between response and explanatory variables. This is especially useful since water
 265 chemistry variables are influenced by each other or by the climatic drivers (von Brömssen et
 266 al., 2021, Ezzati et al., 2023).

267
 268 $F(E(y)) = f(x_1, \dots, x_n) = \beta + s_1(x_1) + s_2(x_2) + \dots + s_n(x_n)$ Eq.

269 2

270

271

272 where y_i is the response variable for $i=1, \dots, n$ and x_1, \dots, x_n are independent variables. $E(y)$ is
 273 the expected value of y [as an alternative for the response distribution apart from the normal
 274 distribution], β is the model intercept, and $s_{i=1-n}(x)$ is a smooth function wrapping the
 275 independent variable.

276

277 The resulting accuracy in capturing the relationships between response variables (water quality
 278 dynamics, i.e. nutrient losses) and explanatory data (climatic drivers of Tair, Tsoil,
 279 Precipitation, and discharge) was checked based on the flexibility of the curve fitting to
 280 multiple variables (Ezzati et al., 2023). A low p-value indicated that residuals were not
 281 randomly distributed (hence there is enough data to capture subtle non-linear linkage between
 282 variables and showing significance of explanatory variables), and a significant downward or
 283 upward trend was identified if the entire confidence band of the derivate of the trend

284 smoothness was below or above zero, respectively (von Brömssen et al., 2021). Here, the
285 “mgcv” package in the R statistical software (Wood, 2017) was applied while “gam.check” (as
286 a tool provided by “mgcv”) was used to check model assumptions and assess effective degrees
287 of freedom for the model terms.

288

289 **3. Results**

290 **3.1. Hydrochemistry**

291 Timoleague and Corduff had the highest annual average water discharge with 643 and 616 mm
292 yr⁻¹, respectively. Cregduff outlet is in an estimated spring contribution zone which explains its
293 very low discharge compared to the rest. However, among the other remaining five catchments,
294 Dunleer had the lowest mean annual discharge (402 mm yr⁻¹) (Fig. 3). The values of the other
295 response variables varied significantly among the six study catchments due to their diverse
296 characteristics and management. Ballycanew had the largest mean SS concentration with a
297 negatively skewed observed concentrations. Cregduff and Corduff had the lowest middle
298 quartile in all water pollutant variables, while the ranges of TP and TRP in Castledockrell were
299 as low as in those two catchments, but the NO₃-N range was the highest among all catchments.
300 Dunleer and Timoleague had normally distributed observed concentrations in NO₃-N and TP
301 and TRP concentrations, and a positively skewed SS with outliers.

302

303 According to the farming calendar and local catchment advisors, slurry spread during
304 spring/summer time is more common in grassland areas where as ploughing and fertiliser
305 application take place during spring on arable land (Fig. 2). For example, in Castledockrell, in
306 which the NO₃-N concentration was highest among all the study catchments (Fig 3), the
307 significant increase of this nutrient in April coincided with chemical fertiliser application and
308 ploughing/sowing (Fig. 2).

309

310

310 **Figure 3**

311

312 **3.2. Monthly trend analysis**

313 The poorly drained catchments of Ballycanew and Corduff and the moderately-drained Dunleer
314 had more frequently occurring monthly trend changes in the concentrations of $\text{NO}_3\text{-N}$ and P
315 compared to the other study catchments (Fig. 4). The increase in monthly trends of nutrient
316 losses in Ballycanew occurred during January, May, and September, while in Dunleer, the
317 increase occurred during October-April. Corduff experienced an increase in monthly trend in
318 nutrient losses throughout the year except in April and December (Fig. 4).

319 The monthly trend analysis of climatic drivers in Ballycanew and Corduff also showed more
320 fluctuations compared to the other study catchments (Fig. 4). The air and soil temperature
321 increased during summer in Ballycanew and Castledockrell. These two catchments are
322 geographically close, yet, the changes in weather pattern were not identical.

323 There was a highly significant increase in TP and TRP concentrations leaving the Timoleague
324 catchment during September; however, it does not coincide with a significant increase in the
325 corresponding monthly trends of any of the climatic variables (Fig. 4).

326 Changes in monthly trends in climatic drivers were observed more frequently in Corduff
327 compared to the other study catchments (Fig. 4). Dunleer experienced significant changes in
328 potential evapotranspiration during September-March. There was also overlaps of the changes
329 in monthly trends in nutrient concentrations with monthly trends in climatic drivers in Dunleer
330 (during January, March, November, December).

331

332 **Figure 4**

333

334 **3.2. Significance of drivers from GAMs**

335 None of the climatic drivers were significant in regulating $\text{NO}_3\text{-N}$ losses during January-March,
336 September, and December. Similarly, none of the drivers were significant in regulating TP/TRP
337 concentration trends during April-May, and October, in any of the study catchments (Table 2
338 and 3). However, May-August and October-January were generally the time windows when
339 significant changes in monthly trends occurred.

340 Discharge appeared to be one of the main regulators of $\text{NO}_3\text{-N}$ losses during months in which,
341 a significant increase of $\text{NO}_3\text{-N}$ concentrations was observed. The significance of discharge
342 occurred alongside its climatic drivers of Tair, Tsoil, and Precipitation. Tair was more

343 associated with increases in NO₃-N and Precipitation with P. Summer time appeared to be the
344 period in which most of the trend changes in the concentrations of both nutrients (Figure 4)
345 and the drivers of concentrations including climatic drivers (Table 2 and 3) took place.

346 In terms of NO₃-N losses, Corduff was mostly influenced by climatic drivers followed by
347 Dunleer, Castledockrell, and Timoleague (Table 2). In terms of P (TP and/or TRP) losses,
348 Dunleer and Corduff, followed by Timoleague and Ballycanew exhibited more susceptibility
349 to changes in weather patterns manifested in the 11 year long time series (Table 3).

350

351

352

Table 2

353

354

Table 3

355

356 4. Discussion

357 Discharge, in relation to its fluctuations with weather changes, was one of the most significant
358 drivers of regulating concentrations of both N and P. Investigation of the dynamics between
359 discharge and nutrient concentrations under changing climate (and different
360 hydrometeorological conditions) is gaining more interest (D'Amario et al., 2021; Pettersen et
361 al., 2021). It is now understood that NO₃-N and TP concentration patterns in relation to
362 discharge tend to vary for different geographical settings and be influenced by other drivers
363 e.g., anthropogenic activities (D'Amario et al., 2021; Vystavna et al., 2023), or changes in
364 weather (Ramos et al., 2022; Ezzati et al., 2023). Studies on the efficiency of management
365 practices under climate change scenarios have shown that depending on the influential
366 catchment characteristics, some watersheds would be more resilient in face of increased
367 intensity of rainfall (Yuan and Koropeckyj-Cox, 2022). Extreme rainfall events and rainfall
368 outside the growing season are known to not only influence agricultural production by
369 prolonging the growing season, but they are also strongly correlated to discharge, N
370 concentration (Øygarden et al., 2014), and P concentration (Ramos et al., 2022). For example
371 a year with normal precipitation has shown strong chemodynamic behaviour in terms of N

372 concentration-discharge while a year with above normal precipitation suggested chemostatic
373 behaviour (Miller and Lyon, 2021). Chemostatic behaviour is hydrologically controlled in the
374 catchment (Pohle et al., 2021), hence, the concentrations remain stable over a wide range of
375 flows in a concentration-discharge (c-q) pattern due to transport limitation (Bieroza et al.,
376 2018). On the contrary, the rate of change in pollutant concentration in a chemodynamic c-q
377 pattern, is larger than the flow change suggesting a source-limitation mechanism (Basu et al.,
378 2010). Precipitation can change the quantity of discharge as well, which would consequently
379 impact the nutrient fluxes (Grusson et al., 2021). This impact would show itself as either a
380 sudden increase in concentrations due to washing out of diffuse sources of nutrients and
381 sediments (Delkash et al., 2018), e.g. increasing the average P loads by 9% (Ockenden et al.,
382 2016), or gradual decreases due to dilution by the water flow (Zhao et al., 2018).

383

384 Significant increases in Tair and Tsoil were observed in both Ballycanew and Castledockrell
385 study catchments during summer months (June, July, August). These drivers suggested to have
386 significant impacts on nutrient concentrations in all catchments. The impact of higher
387 temperature on NO₃-N concentrations in diverse geographical locations has also been observed
388 in other studies (Buciene et al., 2019; Ezzati et al., 2023). Higher temperatures (warmer
389 summer or winter) will create large N pools in soils and increase mineralisation and
390 nitrification rates (Wang et al., 2016) which leads to an increased risk of concentrations at the
391 start of any ensuing wet period (Melander & Jordan, 2021). Rising temperatures would not
392 only increase the risk of droughts and fluctuating quantities of discharge (Vystavna et al.,
393 2023), but will also change soil shear strength (Zhang et al., 2022), which would consequently
394 cause decreases in multi-year flow and mean active water storage capacity in catchments
395 (Vystavna et al., 2023). On the other hand, water limitation in summer droughts affects grass
396 growth by reducing uptake from spring N fertilizers. In order to compensate for that, higher
397 mineral N fertilizers may in some cases be applied (Kundel et al., 2021) which would increase
398 the risk of concentrations during winter time. Also, as a large pool of excessive N has been
399 built up, a higher water discharge will increase losses with the first rain after a prolonged
400 summer period (Ulen et al., 2009; Outram et al., 2014; Mellander and Jordan, 2021). However,
401 if the groundwater has a high storage potential or long residence time (Fenton et al., 2011), or
402 the ditches have low equilibrium phosphorus concentration (Ezzati et al., 2021), the excess N
403 and P, respectively, will be naturally attenuated before leaving the catchment.

404

405 The three groundwater-fed catchments (Castledockrell, Cregduff, and Timoleague with
406 dominant soil types of Al-rich, Al and Ca rich, and Fe-rich, respectively) showed a chemostatic
407 behaviour for P which was not affected by changes in the weather patterns (Mellander et al.,
408 2022^b). Catchment characteristics can significantly change the relationships between discharge
409 and stream water nutrient concentrations by changing the strength and nature of stream flows
410 (Gao et al., 2021) or provide a natural attenuation function due to soil chemistry (Ezzati et al.,
411 2021). The biogeochemical properties of soils in these catchments are also indicative of soil
412 high affinity to bind P as long as the soil has not become P-saturated to act as a P-source rather
413 than a P-sink (Ezzati et al., 2020). In Castledockrell, the significant increase in NO₃-N
414 concentrations in April over the 11 years coincided with chemical fertiliser application and
415 ploughing/sowing (Fig. 2). This catchment is not hydrologically risky (i.e., extreme
416 hydrological events do not create a shift in nutrient dynamics), and was categorised as
417 chemostatic in terms of P (i.e., higher flow will not lead to higher nutrient losses (Bieroza et
418 al., 2018)). According to Mellander & Jordan (2021), the groundwater-fed well drained “N loss
419 risky” Castledockrell catchment, becomes “P loss risky” during winter storms. This transition
420 happens when SMD has been 0 mm for two or more consecutive days, and the effective rain
421 exceeded 15 mm/day (Mellander and Jordan, 2021). Timoleague catchment experienced a
422 significant ($P < 0.001$) increase in P-concentrations during September and a first-level
423 significant increase ($P < 0.1$) in NO₃-N concentrations in January, April, and November, over
424 the 11 year study period. Spring barley, wheat, and maize were the dominant crops, with the
425 latter requiring a large amount of slurry spreading which starts from mid-January and continues
426 until October (inclusive). Chemical fertilisers were applied in April. The maize fields are also
427 closer to the catchment outlet, which may contribute to the high concentrations in the outlet.

428

429 In the Ballycanew catchment, the increases in TP and TRP concentrations occurred during
430 January, May, and September; yet, no particular farming practice could be associated with
431 increases in P flux during January. Grazing (which occurs February-November) coincided with
432 cutting grasses in May. However, slurry spreading in September coincided with a significant
433 increase in the trend of monthly Tair values. Assessment of the high-resolution data revealed
434 that precipitation exceeding 10 mm per day consistently caused a sharp increase in P
435 concentrations. The elevated level of concentration returned to background values after few

436 days when the nutrient sources were exhausted or not hydrologically connected. In addition,
437 continuous rainfall over several days caused an increase in TP concentrations, regardless of the
438 amount of rainfall as more source areas were connected. Remarkably, an extreme hydrological
439 event can turn the P-risky catchment of Ballycanew to N-risky as well. Ireland experienced a
440 series of extreme weather events including a severe summer drought in 2018 (Falzoi et al.,
441 2019). Following the rainfalls in September in the same year, the daily average concentration
442 of NO₃-N reached about 6 times larger than the Environmental quality Standard (EQS) in
443 Ballycanew and it remained high throughout the year (Mellander and Jordan, 2021).

444

445 Corduff exhibited the highest chemodynamic behaviour among the studied catchments.
446 According to Mellander et al. (2018), the average annual runoff coefficients (fraction of rainfall
447 that appears as runoff) between 2009-2015, in descending order, were Corduff (0.57),
448 Timoleague (0.55), Castledockrell (0.54), Dunleer (0.48), and Ballycanew (0.48), (Cregduff is
449 a spring contribution zone). This explains the high chemodynamic behaviour (Bieroza et al.,
450 2018) of Corduff (highly affected by changes in the climate due to hydrological
451 characteristics). The slurry spread during July seemed to have a significant impact on the
452 concentrations of all nutrients. Corduff is mainly permanent grassland with sheep being the
453 main grazing animal from March-November. Increasing trends in P (TP and TRP) and NO₃-N
454 concentrations in Dunleer **were** observed in six months and one month, respectively. This
455 increase in monthly losses occurred almost as frequently as in Corduff, during October-
456 January, and March-April. However, potential evapotranspiration seemed to be the only driver
457 with an increasing trend (i.e. less precipitation), which may lead to slurry spreading as the
458 weather condition is good. According to local knowledge transfer expert, the ploughing/sowing
459 in Dunleer take place during October and November, potatoes are harvested during November,
460 and beets are harvested during November/December. Hence, further farm-scale investigation
461 is required to understand the changes in nutrient concentrations leaving the Dunleer catchment
462 outlet as it is the most complex catchment among all those included herein.

463

464 The results of analysing high temporal resolution historical weather and water quality data of
465 the past 11 years suggested a relationship between increasing trends in climatic drivers and
466 trends in within-stream nutrient concentrations. However, the significance of the correlation
467 between nutrient concentrations and changes in weather patterns was defined by the catchment

468 characteristics including the drainage status and soil chemistry. The present study showed that
469 the non-climatic indices, such as discharge and SS, which are also affected by the changing
470 weather patterns, were significant drivers of nutrient concentrations mostly during the warmer
471 time period of the study years. Hence, current mitigation strategies require explicit
472 consideration of the more frequent occurrence of extreme weather events and their impacts on
473 water pollutant concentrations. It is important to note that there may be inter-seasonal trends
474 that are not identified when assessing inter-annual trends. For example, in the arable
475 Castledockrell catchment, it is recommended to focus on improving N mitigation measures
476 and/or changes in management decisions during April and November. However, in the other
477 grassland catchments of Corduff and Timoleague, the summer period (May-August) plus
478 October-November; and January, April, and November are important, respectively.
479 Meanwhile, an improved focus on P mitigation and management is required in Ballycanew
480 during January, May, and September; in Dunleer during March, July, October-January, and in
481 Timoleague during September and December.

482

483 **5. Conclusions**

484 This study used long-term and high temporal resolution data (10-min) to investigate the N, P,
485 and SS concentration trends in relation to the changing hydro-meteorological patterns in six
486 agricultural catchments during 2010-2021. The results of historical data analysis showed that
487 some of the catchments exhibited changes in the daily average trends of NO₃-N, TP, TRP,
488 and/or SS concentrations with an associated significant increasing trend of climatic drivers, i.e.
489 air temperature, soil temperature, and precipitation across the same months over the 11 years
490 of monitoring period. The soil chemistry and drainage status of the study catchments are the
491 non-climatically related drivers of nutrients concentrations while changes in air and soil
492 temperature, SMD, and precipitation are the main climatic drivers regulating the nutrient
493 concentrations. The extent of the impact of these climatic drivers highly depend on unique
494 catchment characteristics. Discharge had a significant impact on the concentrations of NO₃-N,
495 TP, TRP, and SS; air and soil temperature were significantly correlated to NO₃-N losses; and
496 precipitation was the major contributor to regulating TP and TRP concentrations across the
497 different catchments.

498 The inter-seasonal and inter-annual trends of losses revealed that the climatically driven
499 changes during the same month over the 11 years are manifested differently in different

500 catchments due to their individual hydroclimatological characteristics and management
501 practices. Hence, catchment specific mitigation strategies are required to overcome future
502 challenges to water quality management and the selection of climate-resilient mitigation
503 measures in view of more frequently occurring extreme weather events and projected climate
504 change scenarios.

505 Hence, building upon the results of the current study herein, we recommend that further studies
506 are needed to investigate possible increases of nutrient concentrations into water bodies from
507 agricultural catchments, using the current projected climate change scenarios in Ireland. This
508 would eventually lead to developing future climate-resilient mitigation strategies.

509

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520

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808 **Caption of Figures**

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811 Figure 1. Location of the six catchments monitored by the Agricultural Catchments
812 Programme. The time series plots show daily average nitrate-N ($\text{NO}_3\text{-N}$) and total phosphorus
813 (TP) concentrations (mg L^{-1}) for the time period of 2010-2021. The values on the y-axis
814 represent maximum concentration detected in the catchments.

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816 Figure 2. Summary of the typical farming calendar for each of the six catchments monitored
817 by the Irish Agricultural Catchments Programme. In order to assist visualisation, similar bar
818 colour are used for similar activity across different catchments.

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820 Figure 3. Box and whisker plots of daily concentrations of SS, $\text{NO}_3\text{-N}$, TP and TRP, and the
821 average annual total discharge leaving the catchment outlets for the period 2010– 2021.
822 Cregduff is a karst spring contribution zone and therefore does not have SS data. The line in
823 the boxplots marks the median point of the data. Whiskers show the upper and lower quartiles
824 (75^{th} and 25^{th} percentiles, respectively). Outliers are shown as circles. The dotted horizontal
825 line shows environmental quality standards (EQS) for $\text{NO}_3\text{-N}$ and TRP.

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827 Figure 4. The average monthly trend analysis of $\text{NO}_3\text{-N}$, TP, TRP concentrations, and
828 explanatory variables during 2010-2021 using Mann-Kendall tests. The red colour indicates
829 increasing trends and the green colour indicates decreasing trends. The level of significance is
830 presented by the different shades of any individual colour.

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838 **Caption of Tables**

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840
841 Table 1. Catchments characteristics. Precipitation, soil and air temperatures, potential
842 evapotranspiration (PET), and soil moisture deficit (SMD) values are averaged for the 11 hydrological
843 years (1st October 2010 – 30th September 2021).

844
845 Table 2. The significance of climatic drivers in regulating NO₃-N concentrations with a
846 significant monthly trend during water years spanning 2010-2021. Only catchments and
847 months with changes are included.

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849
850 Table 3. The significance of climatic drivers in regulating TP and/or TRP concentrations with
851 a significant monthly trend during water years spanning 2010-2021. . Only catchments and
852 months with changes are included.

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